

# Tropospheric ozone changes and ozone sensitivity from present-day to future under shared socio-economic pathways

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**Abstract.** Tropospheric ozone is important to future air quality and climate. We investigate ozone changes and ozone sensitivity to changing emissions in the context of climate change from the present day (2004–2014) to the future (2045–2055) under a range of shared socio-economic pathways (SSPs). We apply the United Kingdom Earth System Model, UKESM1, with an extended chemistry scheme including more reactive volatile organic compounds (VOCs) to quantify ozone burdens as well as ozone sensitivities globally and regionally based on nitrogen oxide ( $\text{NO}_x$ ) and VOC [mixing ratio concentrations](#). We show that the tropospheric ozone burden increases by 4 % under a development pathway with higher  $\text{NO}_x$  and VOC emissions (SSP3-7.0), but decreases by 7 % under the same pathway if  $\text{NO}_x$  and VOC emissions are reduced (SSP3-7.0-lowNTCF) and by 5 % if atmospheric methane ( $\text{CH}_4$ ) [mixing ratio concentrations](#) are reduced (SSP3-7.0-lowCH4). Global mean surface ozone [mixing ratio concentrations](#) are reduced by 3–5 ppb under SSP3-7.0-lowNTCF and by 2–3 ppb under SSP3-7.0-lowCH4. However, surface ozone changes vary substantially by season in high-emission regions under future pathways, with decreased ozone [mixing ratio concentrations](#) in summer and increased ozone [mixing ratio concentrations](#) in winter when  $\text{NO}_x$  emissions are reduced. VOC-limited areas are more extensive in winter (7 %) than in summer (3 %) across the globe. North America, Europe and East Asia are the dominant VOC-limited regions in the present day but North America and Europe become more  $\text{NO}_x$ -limited in the future mainly due to reductions in  $\text{NO}_x$  emissions. The impacts of VOC emissions on ozone sensitivity are limited in North America and Europe because reduced anthropogenic VOC emissions are [partly](#) offset by higher biogenic VOC emissions. Ozone sensitivity is not greatly influenced by changing  $\text{CH}_4$  [mixing ratio concentrations](#). South Asia becomes the dominant VOC-limited region under future pathways. We highlight that reductions in  $\text{NO}_x$  emissions are required to transform ozone production from VOC- to  $\text{NO}_x$ -limitation, but that these lead to increased ozone [mixing ratio concentrations](#) in high-emission regions, and hence emission controls on VOC and  $\text{CH}_4$  are also necessary.

## 20 1 Introduction

Ozone ( $\text{O}_3$ ) is a chemically reactive component in the atmosphere that is produced from natural and anthropogenic sources. Emissions of  $\text{O}_3$  precursors including nitrogen oxides ( $\text{NO}_x$ ), volatile organic compounds (VOCs), methane ( $\text{CH}_4$ ) and carbon monoxide (CO) lead to the formation of  $\text{O}_3$  by a series of photochemical reactions in the presence of sunlight.  $\text{O}_3$  has important

impacts on human health, ecosystems and climate change (Lefohn et al., 2018; Zhang et al., 2019; Agathokleous et al., 2020).  
25 O<sub>3</sub> mixing ratio concentrations are largely governed by the magnitudes of O<sub>3</sub> precursor emissions, transport, deposition and  
transport from the stratosphere. O<sub>3</sub> exerts a positive radiative forcing on climate forcing (Stevenson et al., 2013; O'Connor et al.,  
2021; Thornhill et al., 2021a), and changes in climate in turn influence ozone (Fiore et al., 2012; Doherty et al., 2013). Climate  
change can alter natural emissions of biogenic VOCs (BVOC), lightning NO<sub>x</sub> and CH<sub>4</sub>, along with temperature, humidity,  
convection and clouds, which further influence O<sub>3</sub> mixing ratio concentrations (Thornhill et al., 2021b). The interactions  
30 between air quality and climate play an important role in the coupled Earth system, and we focus on the impacts of future  
emissions in the context of climate change on tropospheric O<sub>3</sub> in this study.

The tropospheric O<sub>3</sub> burden is controlled by the amount of O<sub>3</sub> production, O<sub>3</sub> destruction, O<sub>3</sub> deposition and the O<sub>3</sub>  
transport from the stratosphere (Lelieveld and Dentener, 2000; Wild, 2007). From pre-industrial times to the present day, the  
tropospheric O<sub>3</sub> burden has increased from approximately 240 Tg to 350 Tg mainly due to substantial increases in anthro-  
35 pogenic O<sub>3</sub> precursor emissions (Lamarque et al., 2010; Young et al., 2013; Griffiths et al., 2021). However, regional surface  
O<sub>3</sub> changes between the pre-industrial and present day vary substantially due to different regional emission changes (Turnock  
et al., 2020) and to differences in O<sub>3</sub> sensitivity to NO<sub>x</sub> and VOC emissions. In recent decades, there has been a decrease in  
surface O<sub>3</sub> mixing ratio concentrations in North America and Europe due to emission controls (Simon et al., 2015; Colette  
et al., 2016; Tarasick et al., 2019). In contrast, increases in surface O<sub>3</sub> levels are observed in South Asia and East Asia due  
40 to industrialization, urbanization and social development (Hakim et al., 2019; Lu et al., 2020). Furthermore, while emission  
controls have been implemented across industrial regions of China in recent years, these have focused on emissions of NO<sub>x</sub>  
and particulate matter, and have led to increased O<sub>3</sub> pollution in some places (Wang et al., 2017; Silver et al., 2018).

It is important to investigate O<sub>3</sub> sensitivity to understand how O<sub>3</sub> chemical regimes might change in different parts of the  
world, and to guide suitable emission control strategies. ~~However, few studies have focused on O<sub>3</sub> sensitivity from a global~~  
45 ~~perspective, or on how this might change in the future.~~ O<sub>3</sub> sensitivity; is typically characterised by NO<sub>x</sub>- or VOC-limited O<sub>3</sub>  
~~production regime regimes for O<sub>3</sub> production, and this determines the effectiveness of different emission control strategies.~~  
It is dependent on the relative abundances of NO<sub>x</sub> and VOC concentrations (Sillman, 1999), ~~and determines the extent and~~  
~~effectiveness of different emission control strategies; or of their oxidation products, nitric acid (HNO<sub>3</sub>) and hydrogen peroxide~~  
(H<sub>2</sub>O<sub>2</sub>) (Kleinman, 1994; Sillman, 1995). VOC-limited regimes typically occur in highly urbanised regions with high NO<sub>x</sub>  
50 mixing ratio concentrations in which decreases in NO<sub>x</sub> emissions increase O<sub>3</sub> mixing ratio concentrations, and O<sub>3</sub> production  
increases with higher VOC emissions. In contrast, changes in NO<sub>x</sub> mixing ratio concentrations dominate O<sub>3</sub> changes in  
NO<sub>x</sub>-limited regimes such that decreases in NO<sub>x</sub> emissions decrease O<sub>3</sub> mixing ratio concentrations, and O<sub>3</sub> mixing ratios  
concentrations are less sensitive to VOC emissions. ~~O<sub>3</sub> sensitivity can be assessed by quantifying the ratio between NO<sub>x</sub> and~~  
~~VOC concentrations and we apply this approach in this study to present day and future conditions.~~ O<sub>3</sub> sensitivity indicators  
55 such as the ratios of NO<sub>x</sub>/VOCs and HNO<sub>3</sub>/H<sub>2</sub>O<sub>2</sub> allow us to identify O<sub>3</sub> sensitivity regimes relatively easily. However, most  
studies focus on O<sub>3</sub> sensitivity in specific regions and for short time periods (Dunker et al., 2002; Sillman and West, 2009; Ye  
et al., 2016), leading to inconsistency in the critical indicator values that distinguish O<sub>3</sub> sensitivity regimes. To address this, we  
generalise the approach by quantifying O<sub>3</sub> sensitivity using a consistent indicator across the globe. This is the first time that the

full range of surface chemical environments across the globe has been explored with a global chemistry-climate model, as far as we are aware. We quantify O<sub>3</sub> sensitivity based on the ratio of NO<sub>x</sub> and VOC mixing ratios, and investigate how regional O<sub>3</sub> sensitivity might change in the future.

The shared socio-economic pathways (SSPs) are future emission and climate scenarios accounting for future social, economic and environmental developments (O'Neill et al., 2014; van Vuuren et al., 2014). The SSPs represent a range of levels of policy strength (weak, medium and strong) to control emissions of near-term climate forcers (NTCFs) that include tropospheric O<sub>3</sub>, O<sub>3</sub> precursors and aerosols (Rao et al., 2017). Our study is based on simulations using historical and future SSPs emissions and climate undertaken as part of the Aerosol Chemistry Model Intercomparison Project (AerChemMIP; Collins et al., 2017) and the wider Coupled-Model Intercomparison Project Phase 6 (CMIP6; Eyring et al., 2016). The aim of AerChemMIP is to quantify the effects of chemistry and aerosols on air quality and climate in CMIP6 by conducting historical and future experiments using chemistry-climate models with specified climate and emission trajectories.

We examine tropospheric O<sub>3</sub> and surface O<sub>3</sub> sensitivity under present-day (2004–2014) and future conditions (2045–2055). Model development and application are described in Sect. 2 along with descriptions of the emission and climate scenarios used. We compare and evaluate the present-day tropospheric O<sub>3</sub> burden and surface O<sub>3</sub> mixing ratio concentrations with two different chemistry schemes in Sect. 3. We then investigate the seasonal, daytime and nighttime differences in O<sub>3</sub> changes in the future compared to present-day for different regions in Sect. 4. Analysis of O<sub>3</sub> mixing ratio concentrations and production is used to quantify O<sub>3</sub> sensitivity and to explain contrasting regional O<sub>3</sub> changes in Sect. 5. We then show the changes in O<sub>3</sub> sensitivity between different seasons and scenarios in Sect. 6 and present our conclusions in Sect. 7.

## 2 Materials and methods

### 2.1 Model description, development and application

We use version 1 of the United Kingdom Earth System Model, UKESM1 (Sellar et al., 2019), to reproduce present-day (2004–2014) O<sub>3</sub> mixing ratio concentrations and to predict O<sub>3</sub> responses to emissions and climate in the future (2045–2055). UKESM1 consists of a physical climate model, the Hadley Centre Global Environment Model version 3 (HadGEM-GC3.1), configured with the Global Atmosphere 7.1 and Global Land 7.0 (GA7.1/GL7.0) components (Walters et al., 2019) to which Earth System (ES) processes have been coupled (Sellar et al., 2019). Atmospheric composition is modelled using a state-of-the-art chemistry and aerosol module, the United Kingdom Chemistry and Aerosol model (UKCA; Morgenstern et al., 2009; O'Connor et al., 2014). UKCA includes a stratosphere-troposphere gas-phase chemistry scheme (StratTrop; Archibald et al., 2020a) coupled to the aerosol scheme GLOMAP-mode (Mann et al., 2010; Mulcahy et al., 2020). The model resolution is N96L85 in the atmosphere, with 1.875° in longitude by 1.25° in latitude, 85 terrain-following hybrid height layers and a model top at 85 km.

While the UKESM1 configuration for CMIP6 used the UKCA StratTrop mechanism, this study also uses an extended gas-phase chemistry scheme that incorporates more reactive VOC species to permit a more realistic representation of O<sub>3</sub> production in polluted environments. The extended chemistry scheme (denoted as Ext\_StratTrop hereafter) is based on the StratTrop

scheme and includes oxidation of the additional chemical components propene ( $C_3H_6$ ), butane ( $C_4H_{10}$ ) and toluene ( $C_7H_8$ ) to represent alkenes, alkanes and aromatic VOC classes, as described in Liu et al. (2021). The extended chemistry scheme includes 101 species, 244 bimolecular reactions, 26 uni- and termolecular reactions, 70 photolytic reactions, 5 heterogeneous reactions and 3 aqueous phase reactions for the sulphur cycle.

The atmosphere-only configuration of UKESM1 is used with prescribed sea surface temperatures and sea ice to show the transient impacts of emissions under present-day and future climates. These are prescribed using monthly-mean time-evolving fields from the fully coupled UKESM1. Greenhouse gas mixing ratio concentrations are prescribed as in historical and future simulations conducted by UKESM1 as part of CMIP6 (Meinshausen et al., 2017, 2020).

## 2.2 Emissions and experiments

Present-day CMIP6 anthropogenic and biomass burning emissions are taken from Hoesly et al. (2018) and van Marle et al. (2017), respectively. Biogenic VOC emissions are calculated interactively within the iBVOC emissions scheme (Pacifico et al., 2011) in the Joint UK Land Environmental Simulator (JULES) land-surface scheme which is coupled to UKCA. Other aspects of the emissions used here are the same as described in Turnock et al. (2020). Anthropogenic emissions are categorised into five sectors (industry, power plants, transport, residences and agriculture) as inputs to the model, with independent diurnal and vertical emission profiles applied for each sector (Bieser et al., 2011; Mailler et al., 2013; Liu et al., 2021).

Three CMIP6 SSP scenarios are used for future simulations: SSP3-7.0, SSP3-7.0-lowNTCF and SSP3-7.0-lowCH4. SSP3-7.0 pathway has a ‘regional rivalry’ policy scenario with a large anthropogenic climate forcing signal (a radiative forcing of  $7.0 \text{ W m}^{-2}$  at 2100), ~~and it has~~ weak emission controls on  $O_3$  precursors and aerosols, and rapidly increasing  $CH_4$  mixing ratio concentrations (Fujimori et al., 2017; Rao et al., 2017). SSP3-7.0-lowNTCF and SSP3-7.0-lowCH4 are additional pathways which use the same underlying climate policies as SSP3-7.0. SSP3-7.0-lowNTCF has strong controls on all NTCF emissions. SSP3-7.0-lowCH4 follows SSP3-7.0 but assumes strong mitigation of  $CH_4$  emissions in the future, with 24 % decreases in surface  $CH_4$  mixing ratios from 1802 ppb to 1364 ppb. The last time that historical surface  $CH_4$  mixing ratios were this low was more than 50 years ago, in the late 1960s (Prather et al., 2014). BVOC emissions increase under all these pathways due to a warmer climate. We perform four model experiments in this study to investigate tropospheric  $O_3$  for the present-day (2004–2014) and three future pathways (SSP3-7.0, SSP3-7.0-lowNTCF and SSP3-7.0-lowCH4; 2045–2055). Table 1 shows the model configuration for the four simulations. Table 2 lists the CMIP6 global mean NTCF total surface emissions and surface  $CH_4$  mixing ratio concentrations for the four scenarios.

## 2.3 $O_3$ sensitivity indicators

A number of different indicators have been used to distinguish  $O_3$  sensitivity regimes, and typical indicators are the ratios of  $NO_x$ /VOC mixing ratios or emissions and the ratio of  $HNO_3/H_2O_2$  mixing ratios (Kleinman, 1994; Sillman, 1999). For the  $NO_x$ /VOC ratio, it is often more appropriate to use mixing ratios than emissions because this accounts for emissions, transport, chemical reactions and deposition. Indicators based on  $HNO_3/H_2O_2$  ratios also account for differences in photochemical conditions and VOC reactivity. In a previous study we found that  $O_3$  sensitivity regimes diagnosed with  $HNO_3/H_2O_2$  and

**Table 1.** Model configurations for present-day and future simulations. “Emissions” refers to emissions of O<sub>3</sub> precursors and aerosols. “CH<sub>4</sub> ~~emissions~~” refers to prescribed surface CH<sub>4</sub> ~~mixing ratios~~ concentrations. “SST/SI” refers to prescribed sea surface temperature and sea ice concentrations. “Historical” means that the emissions, CH<sub>4</sub> ~~mixing ratios~~ or SST/SI evolve as for the CMIP6 historical simulations, and “Reference” means that they evolve as for SSP3-7.0. “Low” emissions or CH<sub>4</sub> ~~mixing ratios~~ evolve following SSP3-7.0 but with lower emissions or CH<sub>4</sub> ~~mixing ratios~~.

Experiment name	Time period	Emissions	CH <sub>4</sub> <del>emissions</del>	SST/SI
Present day	2004–2014	Historical	Historical	Historical
SSP370	2045–2055	Reference	Reference	Reference
SSP370_lowNTCF	2045–2055	Low	Reference	Reference
SSP370_lowCH4	2045–2055	Reference	Low	Reference

125 NO<sub>x</sub>/VOC ratios were similar (Liu et al., 2021). However, the HNO<sub>3</sub>/H<sub>2</sub>O<sub>2</sub> indicator is more sensitive to uncertainties in chemical mechanism, and studies have shown that there are errors in the simulation of short-lived radicals in polluted areas (Whalley et al., 2021). The HNO<sub>3</sub>/H<sub>2</sub>O<sub>2</sub> ratio also does not account for gas-aerosol conversion as a termination route for NO<sub>x</sub>, a mechanism that is included in many chemistry-climate models. We hence choose the ratio between NO<sub>x</sub> and VOC mixing ratios as a simple indicator of O<sub>3</sub> sensitivity indicator in this study.

130 We quantify the sensitivity of O<sub>3</sub> to NO<sub>x</sub> and VOC mixing ratios by examining monthly mean O<sub>3</sub> mixing ratios and O<sub>3</sub> net production in each UKESM1 surface grid cell in each of the scenarios in turn. This provides a global overview of the dependence of O<sub>3</sub> and its production on NO<sub>x</sub> and VOC across different environments. It also allows us to determine a globally-averaged critical threshold value distinguishing NO<sub>x</sub>-limited and VOC-limited regimes.

135 We note that dilution of short-lived NO<sub>x</sub> over coarse resolution model grid cells may lead to the underestimation of local NO<sub>x</sub> mixing ratios in high-emission regions. This results in underestimation of NO<sub>x</sub>/VOC ratios in these conditions and the regimes may thus be more VOC-limited in reality than are able to simulate in a global model.

### 3 Model evaluation of tropospheric and surface O<sub>3</sub>

#### 3.1 Comparison of StratTrop and extended chemistry schemes

We first compare averaged tropospheric O<sub>3</sub> burdens, chemical lifetime, chemical production, chemical loss and deposition during 2004–2014 from the extended chemistry scheme (Ext\_StratTrop) with those from the StratTrop chemistry scheme used in AerChemMIP simulations (Table 3). We define the O<sub>3</sub> production rate as the sum of reactions fluxes through HO<sub>2</sub>/RO<sub>2</sub> + NO, and the O<sub>3</sub> loss rate as the sum of O(<sup>1</sup>D) + H<sub>2</sub>O, O<sub>3</sub> + HO<sub>2</sub>/OH/alkenes. The O<sub>3</sub> burden with the extended chemistry scheme (376 Tg) lies at the upper end of the uncertainty range for the observed burden, 340 ± 34 Tg for 2000 (Archibald et al., 2020b). The magnitude of the O<sub>3</sub> burden is also consistent with the CMIP6 multi-model mean burden of 356 ± 31 Tg for 2005–2014 (Griffiths et al., 2021). The extended chemistry scheme produces a 5 % higher tropospheric O<sub>3</sub> burden than

**Table 2.** Overview of global annual mean time-varying surface emissions of NO<sub>x</sub>, VOCs, CO, sulfur dioxide (SO<sub>2</sub>), black carbon (BC) and organic carbon (OC) from anthropogenic (ANT), biomass burning (BB), biogenic (BIO) sources for the present day (2004–2014) and future (2045–2055) SSP3-7.0, SSP3-7.0-lowNTCF and SSP3-7.0-lowCH4. Annual mean surface CH<sub>4</sub> mixing ratio concentrations (ppb) are also shown.

Emission (Tg yr <sup>-1</sup> )		Present day	SSP3-7.0	SSP3-7.0-lowNTCF	SSP3-7.0-lowCH4
NO <sub>x</sub>	ANT	136.0	149.6	68.8	149.6
	BB	13.6	12.1	10.2	12.1
	Total	149.6	161.7	79.0	161.7
VOCs	ANT	156.3	195.3	117.6	195.3
	BB	62.7	57.1	47.5	57.1
	BIO	727.9	786.1	795.2	785.7
	Total	946.9	1038.5	960.3	1038.1
CO	ANT	600.8	662.7	328.4	662.7
	BB	324.6	318.5	264.5	318.5
	Total	925.4	981.2	592.9	981.2
SO <sub>2</sub>	ANT	115.4	95.7	43.0	95.7
	BB	2.1	2.2	1.8	2.2
	Total	117.5	97.9	44.8	97.9
BC	ANT	7.4	9.1	4.4	9.1
	BB	1.7	1.7	1.4	1.7
	Total	9.1	10.8	5.8	10.8
OC	ANT	18.0	23.3	10.1	23.3
	BB	15.0	14.5	11.9	14.5
	Total	33.0	37.8	22.0	37.8
CH <sub>4</sub> (ppb)		1802.8	2471.9	2471.9	1363.7

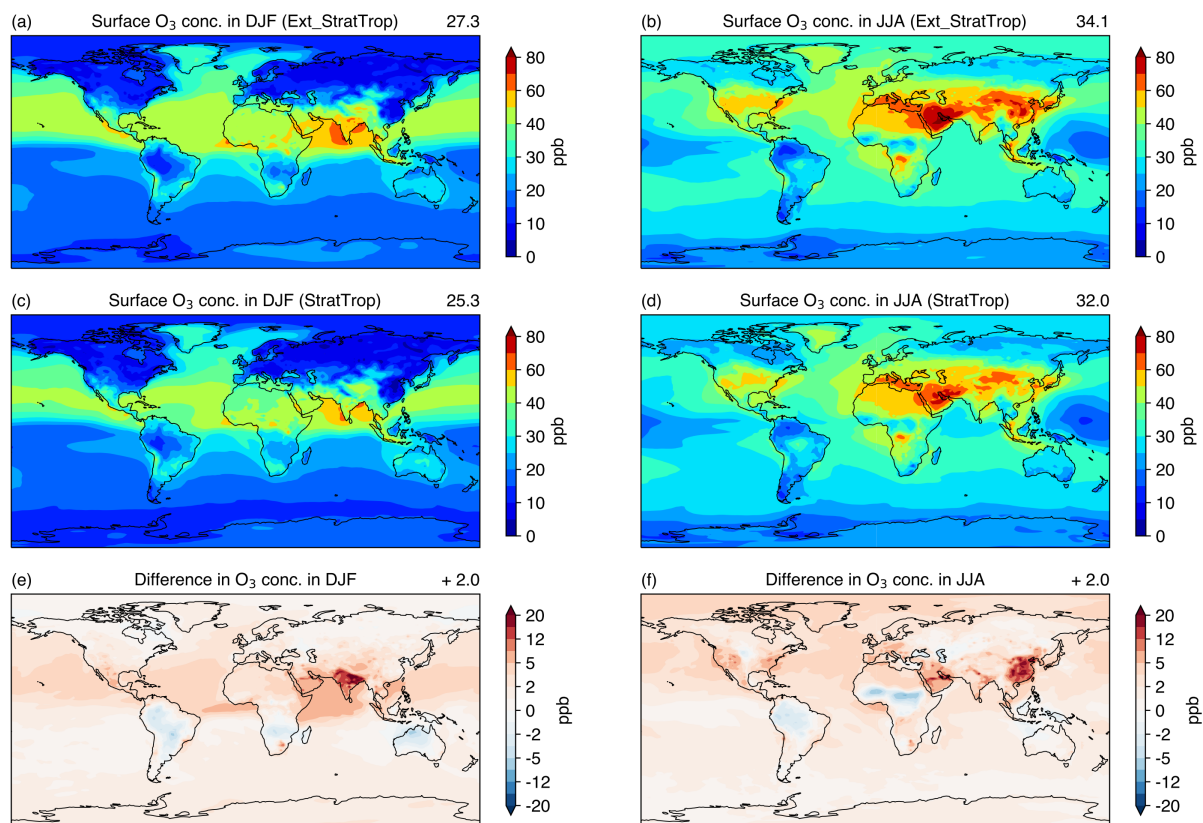
that of StratTrop (358 Tg), demonstrating a more reactive environment for net O<sub>3</sub> formation throughout the troposphere due to reactive VOCs. This is also reflected in the higher rates of chemical O<sub>3</sub> production (11 %), loss (6 %) and deposition (6 %) with the extended chemistry scheme. However, the higher O<sub>3</sub> production is offset by greater O<sub>3</sub> destruction and by faster O<sub>3</sub> deposition, and hence the mean O<sub>3</sub> chemical lifetime remains very similar at about 22 days, which is consistent with previous multi-model estimates of the mean lifetime of 22.2 ± 2.2 days (Stevenson et al., 2006).

Simulated surface O<sub>3</sub> mixing ratio concentrations during 2004–2014 with the two chemistry schemes are compared in Fig. 1. Using the extended chemistry scheme, the spatial distribution of surface O<sub>3</sub> is similar to that using StratTrop in both winter

**Table 3.** Comparison of tropospheric O<sub>3</sub> burden and budget terms during 2004–2014 simulated with StratTrop and Ext\_StratTrop chemistry schemes in UKCA. One standard deviation in the annual terms over the 2004–2014 period is shown. We define the tropopause based on the highest layer with an O<sub>3</sub> ~~mixing ratio~~ concentration less than 150 ppb.

	StratTrop	Ext_StratTrop
O <sub>3</sub> burden (Tg)	358±3	376±3
O <sub>3</sub> lifetime (days)	22.6±0.2	22.5±0.2
O <sub>3</sub> net production (Tg year <sup>-1</sup> )	895±45	996±40
O <sub>3</sub> production (Tg year <sup>-1</sup> )	5698±40	6080±66
O <sub>3</sub> loss (Tg year <sup>-1</sup> )	4803±45	5084±59
O <sub>3</sub> deposition (Tg year <sup>-1</sup> )	883±9	936±9

and summer, but shows a general increase in global O<sub>3</sub> levels of about 2 ppb (Fig. 1a–d) due to inclusion of the additional reactive VOCs. The O<sub>3</sub> increases (Fig. 1e–f) are most notable for South Asia and East Asia due to the relatively high VOC emissions in these regions (Janssens-Maenhout et al., 2015; Huang et al., 2017; Feng et al., 2020). There is a much larger O<sub>3</sub> increase in South Asia in winter than in summer (Fig. 1e), mainly due to greater transport of O<sub>3</sub> precursors during the summer monsoon in South Asia (Lu et al., 2018). This leads to higher O<sub>3</sub> ~~mixing ratio~~ concentrations in winter than that in summer, consistent with Gao et al. (2020), and demonstrates a larger seasonal variation in O<sub>3</sub> ~~mixing ratio~~ concentrations with the Ext\_StratTrop chemistry scheme than with the StratTrop chemistry scheme. In addition, there are substantial O<sub>3</sub> increases in East Asia and in other polluted continents in summer (Fig. 1f) when using the extended chemistry scheme. This is not seen in winter (Fig. 1e) because titration of O<sub>3</sub> by nitric oxide (NO) remains strong, despite high VOC emissions. The relatively small influence of the additional VOCs on O<sub>3</sub> in winter is also seen in other heavily populated regions that have high NO<sub>x</sub> emissions such as North America and Europe.



**Figure 1.** Comparison of present-day seasonal mean surface O<sub>3</sub> mixing ratio concentrations (2004–2014) between the Ext\_StratTrop chemistry scheme (a, b) and the StratTrop chemistry scheme (c, d). Winter time (December–January–February, DJF) and summer time (June–July–August, JJA) global mean O<sub>3</sub> mixing ratio concentrations are shown. Seasonal differences in surface O<sub>3</sub> mixing ratio concentrations between the two chemistry schemes are shown in the bottom panels (e, f).

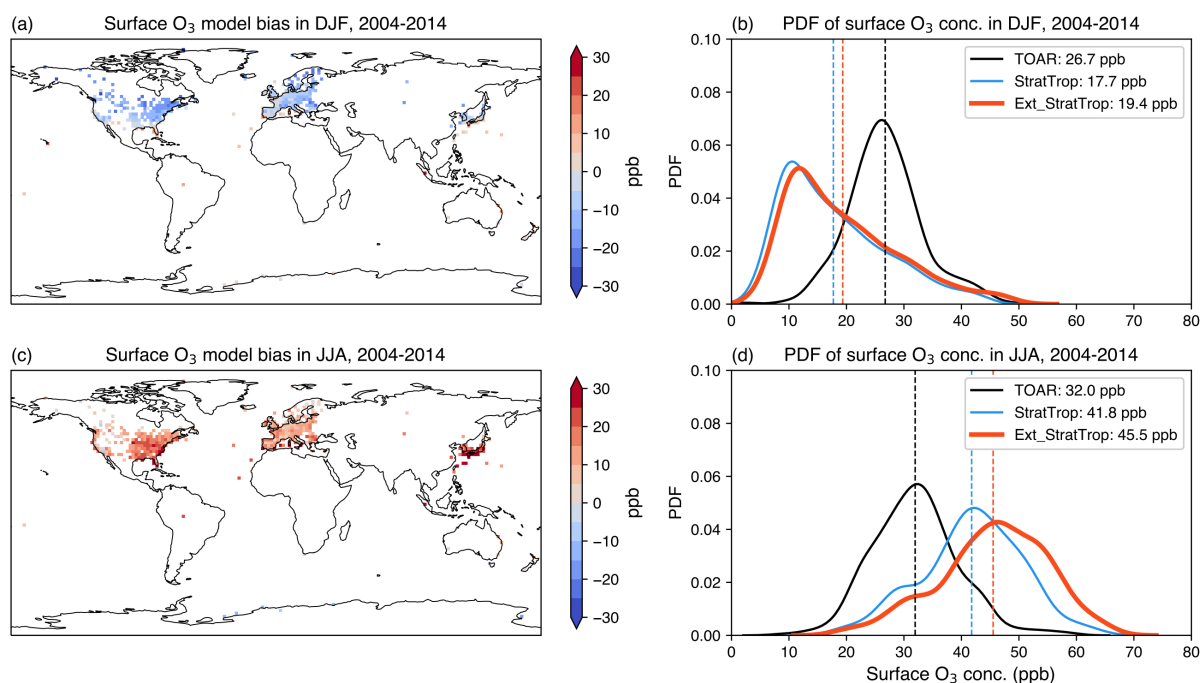
### 3.2 Evaluation of surface O<sub>3</sub> mixing ratio concentrations

165 We now evaluate surface O<sub>3</sub> mixing ratio concentrations simulated with the Ext\_StratTrop chemistry scheme against gridded monthly mean rural observations from the TOAR dataset over the 2004–2014 period (Schultz et al., 2017a, b). Surface O<sub>3</sub> mixing ratio concentrations for winter (DJF) and summer (JJA) during 2004–2014 are shown in Fig. 2. We find that global mean surface O<sub>3</sub> mixing ratio concentrations are underestimated in winter (-7.3 ppb) and overestimated in summer (+13.5 ppb), with relatively small biases in spring and autumn. The positive biases in summer and negative biases in winter for O<sub>3</sub> mixing ratio concentrations are also seen in results using the StratTrop chemistry scheme (Archibald et al., 2020a; Turnock et al., 2020). The model seasonality with both schemes (19.4–45.5 ppb; DJF–JJA) is rather stronger than that observed (26.7–32 ppb; DJF–JJA), but the Ext\_StratTrop chemistry scheme improves the model performance slightly in DJF for surface O<sub>3</sub> (Fig. 2b) despite larger biases in summer (Fig. 2d). Numerical diffusion of O<sub>3</sub> precursor emissions due to coarse model horizontal

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175 resolution may explain the biases (Wild and Prather, 2006; Stock et al., 2014; Fenech et al., 2018), and we note that these  
 180 seasonal  $O_3$  biases are also evident in other chemistry-climate models (Young et al., 2018; Turnock et al., 2020). Insufficient  
 turbulent mixing in the planetary boundary layer may also contribute to the bias (O'Connor et al., 2014), as accumulation of  
 $NO_x$  at the surface leads to greater  $O_3$  production in summer and greater titration by  $NO$  in winter. However, we note that a  
 much more comprehensive chemistry scheme applied in UKCA, the Common Representative Intermediates Mechanism (CRI-  
 Strat), shows similar systematic biases in surface  $O_3$  mixing ratio concentrations, -4.6 ppb in DJF and +12.0 ppb in JJA for  
 the 2010–2018 period (Archer-Nicholls et al., 2020). This suggests that other biases in the model are primarily responsible for  
 the biases in surface  $O_3$ . We choose to apply the extended chemistry scheme as it permits representation of a more appropriate  
 chemical environment for  $O_3$  production in high-emission areas and is thus more suitable to investigate  $O_3$  sensitivity.



**Figure 2.** Modelled surface  $O_3$  biases with the Ext\_StratTrop chemistry scheme for winter (DJF) and summer (JJA) over 2004–2014 (a, c). Probability distribution function (PDF) of seasonal mean  $O_3$  mixing ratio concentrations between observations, StratTrop and Ext\_StratTrop chemistry schemes (b, d). Vertical dashed lines indicate seasonal mean surface  $O_3$  mixing ratio concentrations. Observations from the Tropospheric Ozone Assessment Report (TOAR) dataset (Schultz et al., 2017a) are used for comparison.

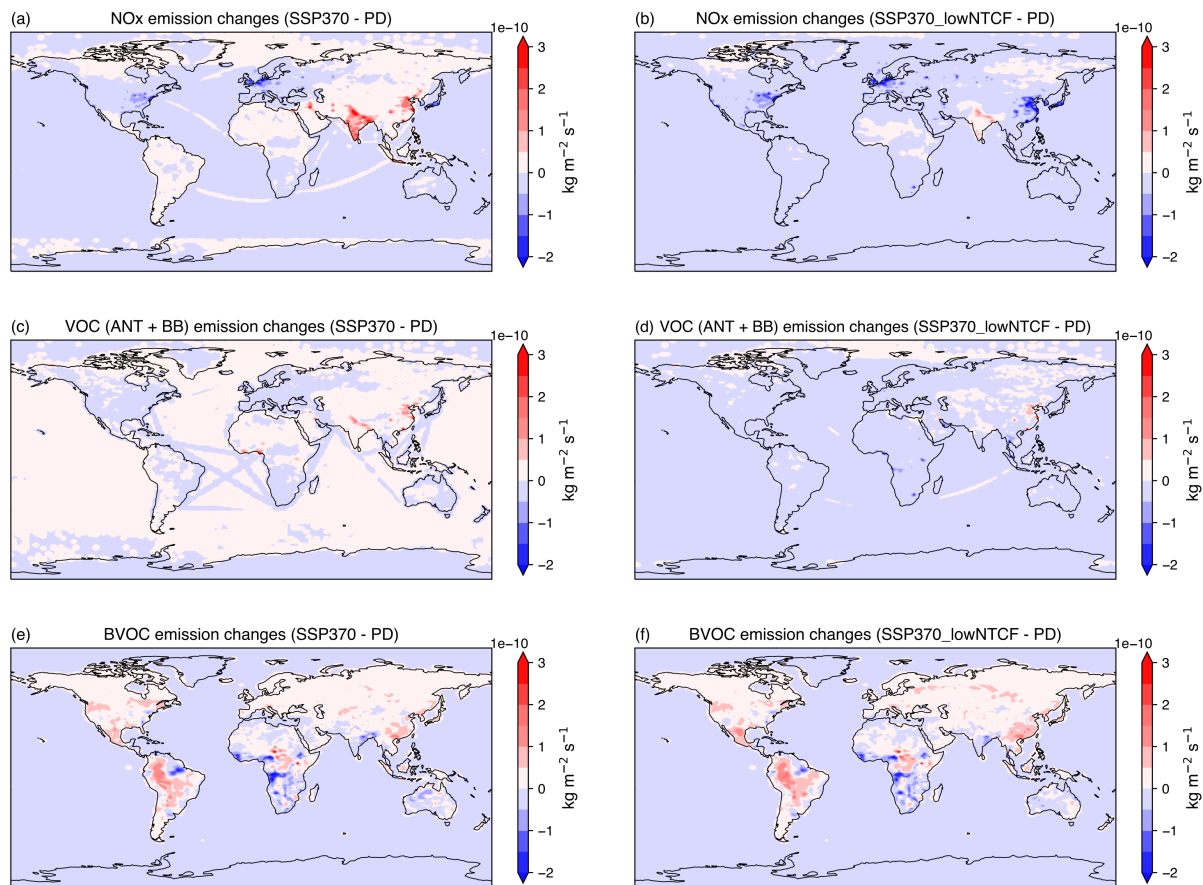
## 4 O<sub>3</sub> changes under future scenarios

### 4.1 Emission changes

185 The SSP3-7.0 pathway is characterized by relatively strong emission controls in some parts of the world such as North America and Europe but weaker controls or emission increases elsewhere. Increases in NO<sub>x</sub> and VOC emissions from anthropogenic and biomass burning sources are seen in Central and South America, North Africa, the Middle East, and South and East Asia (Fig. 3a, c). SSP3-7.0-lowCH<sub>4</sub> has the same NTCF emissions as SSP3-7.0 but lower CH<sub>4</sub> emissions that lead to lower CH<sub>4</sub> mixing ratio concentrations (Table 2). SSP3-7.0-lowNTCF represents strong emission controls across the globe, with  
190 reductions in emissions in most major high-emission regions except for South Asia (Fig. 3b, d). Total BVOC emission changes are driven by changes in land-use, vegetation and temperature. Fig. 3e, f shows general increases in total BVOC emissions in most parts of the Northern Hemisphere except for South Asia in the future, which partly offset the decreased anthropogenic and biomass burning VOC emissions in North America and Europe. Similar BVOC emission changes are found under SSP3-7.0 and SSP3-7.0-lowNTCF relative to the present day as the climate change signal and carbon dioxide concentrations under future  
195 pathways are the same.

### 4.2 Tropospheric O<sub>3</sub> changes

Changes in tropospheric O<sub>3</sub> from the present day to the future are shown in Table 4. Changes in NTCF emissions control O<sub>3</sub> burden changes with a 4 % increase in the O<sub>3</sub> burden under SSP3-7.0 and a 7 % decrease under SSP3-7.0-lowNTCF relative to the present day. Changes in O<sub>3</sub> production rates are also controlled by changes in NTCF emissions, and there is higher  
200 O<sub>3</sub> production under SSP3-7.0 and lower O<sub>3</sub> production under SSP3-7.0-lowNTCF. The decrease in the O<sub>3</sub> burden (5 %) under SSP3-7.0-lowCH<sub>4</sub> is slightly less than that under SSP3-7.0-lowNTCF (7 %), and shows that reductions in CH<sub>4</sub> mixing ratio concentrations effectively reduce the tropospheric O<sub>3</sub> burden despite high NTCF emissions under SSP3-7.0-lowCH<sub>4</sub>. O<sub>3</sub> production rates under SSP3-7.0-lowCH<sub>4</sub> are slightly higher than in the present day partly due to higher hydroxyl radical (OH) concentrations that promote O<sub>3</sub> production. However, these higher O<sub>3</sub> production rates are offset by higher O<sub>3</sub> loss rates, and  
205 result in lower O<sub>3</sub> net production under SSP3-7.0 and SSP3-7.0-lowCH<sub>4</sub>. We find that the O<sub>3</sub> chemical lifetime decreases slightly by 0.4–1.6 days under future pathways partly due to decreased O<sub>3</sub> net production, and partly due to increased O<sub>3</sub> loss associated with higher temperature and humidity in a warmer climate (Young et al., 2018). Changes in O<sub>3</sub> dry deposition rates principally reflect changes in surface O<sub>3</sub> mixing ratio concentrations, although high temperatures under a warmer climate may reduce O<sub>3</sub> deposition rates due to vegetation stress (Lin et al., 2020).



**Figure 3.** Differences in annual mean surface emissions of  $\text{NO}_x$  (**a**, **b**), anthropogenic and biomass burning VOCs (**c**, **d**) and total biogenic VOCs (**e**, **f**) between the present day (PD; 2004–2014) and SSP3-7.0 and SSP3-7.0-lowNTCF (2045–2055).

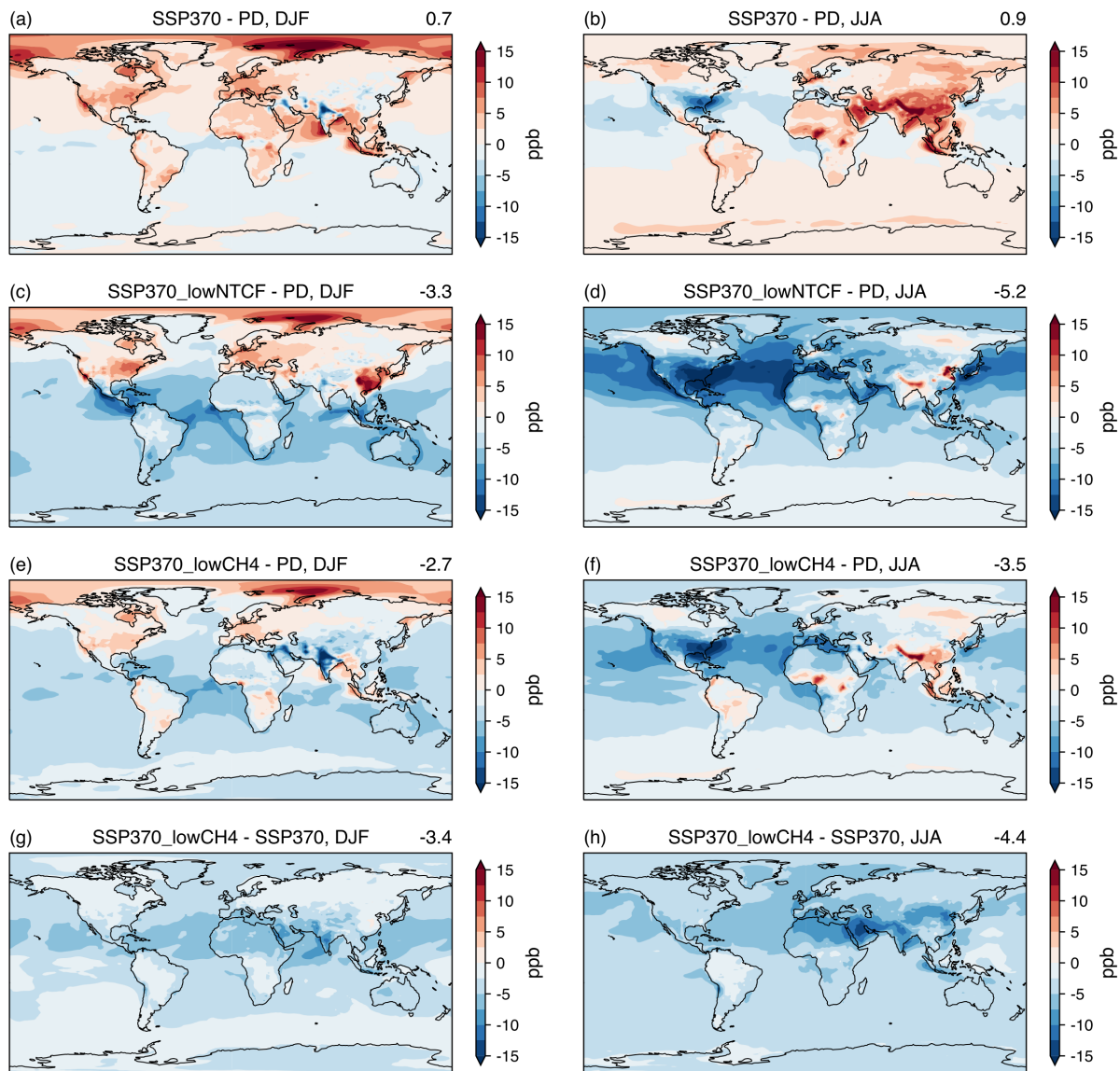
**Table 4.** Comparison of the tropospheric  $\text{O}_3$  budget under present-day (2004–2014) and future conditions (2045–2055).

	Present day	SSP3-7.0	SSP3-7.0-lowNTCF	SSP3-7.0-lowCH4
$\text{O}_3$ burden (Tg)	$376 \pm 3$	$393 \pm 5$	$351 \pm 7$	$357 \pm 6$
$\text{O}_3$ lifetime (days)	$22.5 \pm 0.2$	$20.9 \pm 0.1$	$22.1 \pm 0.1$	$21.1 \pm 0.2$
$\text{O}_3$ net production ( $\text{Tg year}^{-1}$ )	$996 \pm 40$	$934 \pm 38$	$638 \pm 60$	$869 \pm 32$
$\text{O}_3$ production ( $\text{Tg year}^{-1}$ )	$6080 \pm 66$	$6728 \pm 43$	$5536 \pm 108$	$6108 \pm 65$
$\text{O}_3$ loss ( $\text{Tg year}^{-1}$ )	$5084 \pm 59$	$5795 \pm 45$	$4898 \pm 78$	$5239 \pm 56$
$\text{O}_3$ dry deposition ( $\text{Tg year}^{-1}$ )	$936 \pm 9$	$961 \pm 9$	$810 \pm 19$	$856 \pm 14$

### 210 4.3 Surface seasonal O<sub>3</sub> changes

Seasonal differences in simulated surface O<sub>3</sub> mixing ratios between the present day (2004–2014) and future pathways (2045–2055) are shown in Fig. 4a–f, along with a comparison between SSP3-7.0 and SSP3-7.0-lowCH<sub>4</sub> (Fig. 4g, h). SSP3-7.0 represents less stringent emission control policies, and has slightly higher global mean O<sub>3</sub> mixing ratios (0.7–0.9 ppb; Fig. 4a, b) than the present day. In contrast, tightened emission controls under SSP3-7.0-lowNTCF reduce surface O<sub>3</sub> mixing ratios substantially across many parts of the world (3.3–5.2 ppb; Fig. 4c, d). The reduction in CH<sub>4</sub> mixing ratios from 1803 to 1364 ppb under SSP3-7.0-lowCH<sub>4</sub> relative to the present day successfully reduces O<sub>3</sub> mixing ratios (2.7–3.5 ppb; Fig. 4e, f) for regions that show O<sub>3</sub> increases under SSP3-7.0. SSP3-7.0-lowCH<sub>4</sub> with reduced CH<sub>4</sub> mixing ratios alone shows uniform O<sub>3</sub> decreases across the globe (3.4–4.4 ppb; Fig. 4g, h) compared with SSP3-7.0, where mean CH<sub>4</sub> mixing ratios are 2472 ppb, and this offsets high O<sub>3</sub> levels in regions with high NTCF emissions. This demonstrates the importance of CH<sub>4</sub> in governing surface O<sub>3</sub> mixing ratios, and the need to account for CH<sub>4</sub> in mitigating O<sub>3</sub> pollution in future (Fiore et al., 2008; Allen et al., 2021). We highlight that O<sub>3</sub> changes vary by season. From Fig. 4a, c we can see that O<sub>3</sub> mixing ratios generally increase in winter in continental areas such as North America and Europe under SSP3-7.0 and SSP3-7.0-lowNTCF. These regions have large reductions in NTCF emissions under future pathways, and thus there is less O<sub>3</sub> titration due to lower NO<sub>x</sub> emissions in these regions. Since NO<sub>x</sub> emissions decrease in East Asia under SSP3-7.0-lowNTCF, we also see O<sub>3</sub> increases in winter in this region. This highlights that NO<sub>x</sub> emission reductions are not beneficial for reducing surface O<sub>3</sub> mixing ratios in winter. Conversely, NO<sub>x</sub> emission increases under SSP3-7.0 lead to O<sub>3</sub> decreases in winter in South and East Asia for the same reason. The situation is different in summer, with O<sub>3</sub> decreases in North America and Europe, but O<sub>3</sub> increases in South and East Asia under SSP3-7.0 (Fig. 4b). This demonstrates a shift in O<sub>3</sub> sensitivity from VOC limitation in winter to NO<sub>x</sub> limitation in summer. O<sub>3</sub> changes in summer and winter are generally consistent in South America and Africa, and reflect NO<sub>x</sub> limitation in these regions throughout the year. For regions that are projected to have lower NO<sub>x</sub> emissions in SSP3-7.0-lowNTCF, such as eastern China, we note that O<sub>3</sub> mixing ratios increase in both winter and summer. The industrial regions of China are in VOC limited regimes throughout the year, and thus decreased NO<sub>x</sub> emissions increase O<sub>3</sub> mixing ratios (Jin and Holloway, 2015; Wang et al., 2021; Liu et al., 2021). This suggests that reductions in both NO<sub>x</sub> and VOC emissions may be needed to reduce O<sub>3</sub> in these regions.

Surface O<sub>3</sub> mixing ratios are also influenced by climate change, reflecting changing natural emissions, O<sub>3</sub> production and destruction rates and O<sub>3</sub> deposition rates (Doherty et al., 2013, 2017). Global annual mean surface O<sub>3</sub> mixing ratios decrease by 1 ppb with a 1.5 °C temperature rise, but show little change in continental areas (Naik et al., 2021). This is principally due to increased humidity and greater O<sub>3</sub> destruction in oceanic areas, but in continental areas these effects may be offset by O<sub>3</sub> increases due to higher soil NO<sub>x</sub> (Romer et al., 2018) and BVOC emissions, and by decreased O<sub>3</sub> deposition rates (Lin et al., 2020). O<sub>3</sub> mixing ratios can also be impacted by more frequent and intense heat waves under a warmer climate (Schnell and Prather, 2017; Ma et al., 2019). We find that the resulting changes in surface O<sub>3</sub> mixing ratios in continental regions due to

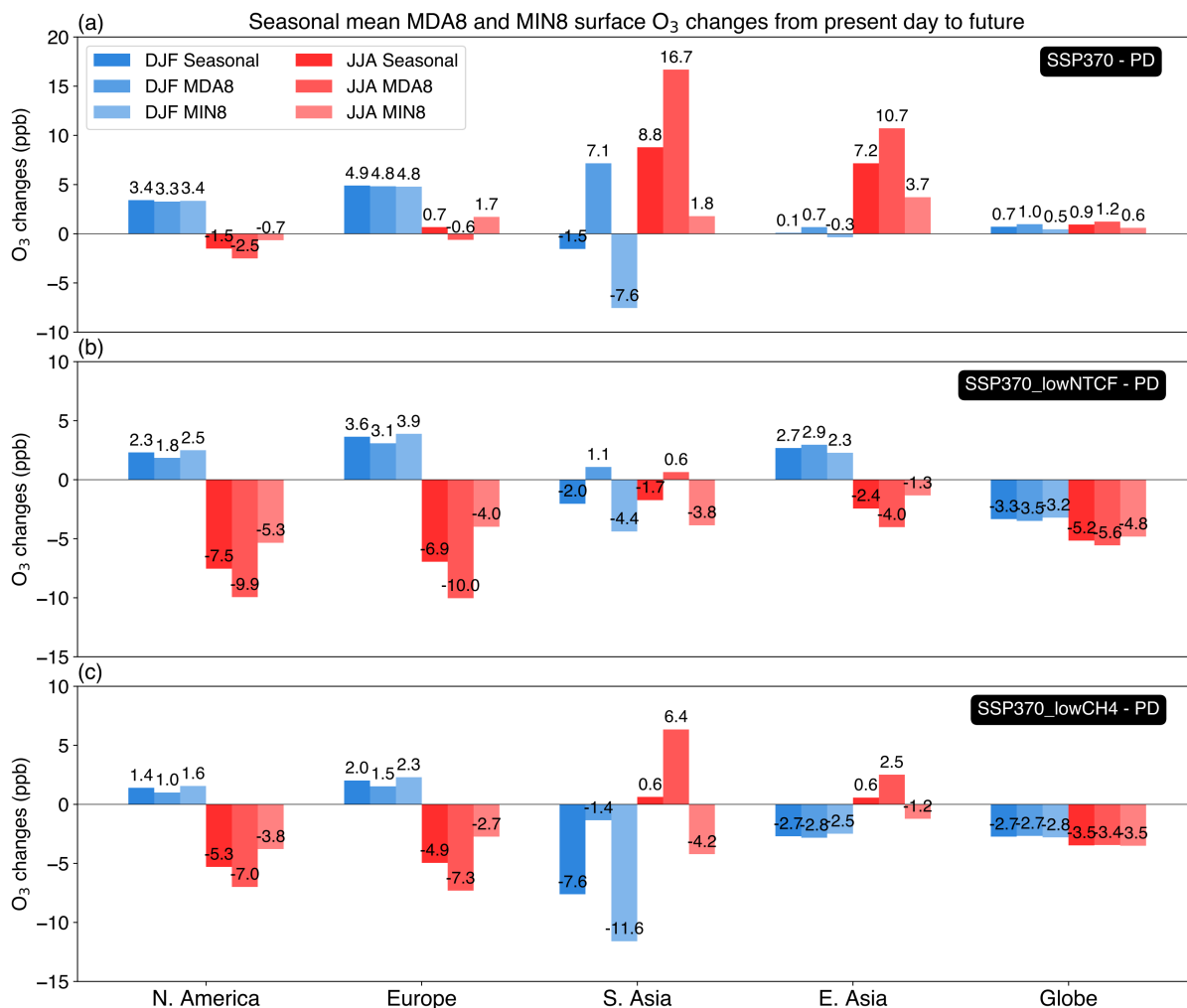


**Figure 4.** Differences in seasonal mean surface O<sub>3</sub> mixing ratios between the present day (PD; 2004–2014) and the future SSP3-7.0 (a, b), SSP3-7.0-lowNTCF (c, d) and SSP3-7.0-lowCH4 (e, f) (2045–2055) in winter (DJF) and summer (JJA). Differences between SSP3-7.0 and SSP3-7.0-lowCH4 are shown (g, f) to isolate the impacts of reduced CH<sub>4</sub> mixing ratios. Absolute global mean O<sub>3</sub> changes (ppb) are shown at the right top of each panel.

climate change are relatively small, and reduction in anthropogenic emissions is the dominant factor governing surface O<sub>3</sub> mixing ratios in the near future.

245 We also examine daytime and nighttime O<sub>3</sub> changes for different regions under future pathways because daytime O<sub>3</sub> mixing ratioconcentrations are typically more relevant for human health. For daytime O<sub>3</sub> concentrations we consider the maximum daily average 8 h O<sub>3</sub> mixing ratioconcentration (MDA8), an important metric used to evaluate O<sub>3</sub> impacts on human health. Nighttime O<sub>3</sub> mixing ratioconcentrations are correspondingly given by the minimum daily average 8 h O<sub>3</sub> mixing ratioconcentration (MIN8). We show a comparison of the changes in these metrics in Fig. 5.

250 In general, differences in global mean O<sub>3</sub> changes are relatively small in both daytime and nighttime under all pathways. However, in South and East Asia with increased NO<sub>x</sub> emissions under SSP3-7.0 and SSP3-7.0-lowCH4, O<sub>3</sub> mixing ratiosconcentrations tend to increase in daytime but show smaller increases or reduction at night. This demonstrates the impact of O<sub>3</sub> titration by NO at nighttime in high-NO<sub>x</sub> environments. In North America and Europe that have lower NO<sub>x</sub> emissions in the future, daytime O<sub>3</sub> mixing ratioconcentrations decrease greatly in summer, but daytime and nighttime changes are  
255 similar in winter, demonstrating the large influence of NO<sub>x</sub> emissions on summer daytime O<sub>3</sub> mixing ratioconcentrations. The substantial differences between daytime and nighttime O<sub>3</sub> changes suggest that the underlying impacts of O<sub>3</sub> changes on human health are likely to be larger than those estimated using seasonal mean O<sub>3</sub> mixing ratioconcentrations, and this is particularly important for high-emission areas with increased NO<sub>x</sub> emissions in the future.



**Figure 5.** Seasonal mean maximum daily average 8 h (MDA8) and minimum daily average 8 h (MIN8) surface O<sub>3</sub> changes for North America, Europe, South Asia, East Asia and the globe from the present day (PD; 2004–2014) to SSP3-7.0 (a), SSP3-7.0-lowNTCF (b) and SSP3-7.0-lowCH4 (c) (2045–2055). DJF and JJA situations are shown with blue and red bars, respectively.

## 5 O<sub>3</sub> sensitivity in the present day and the future

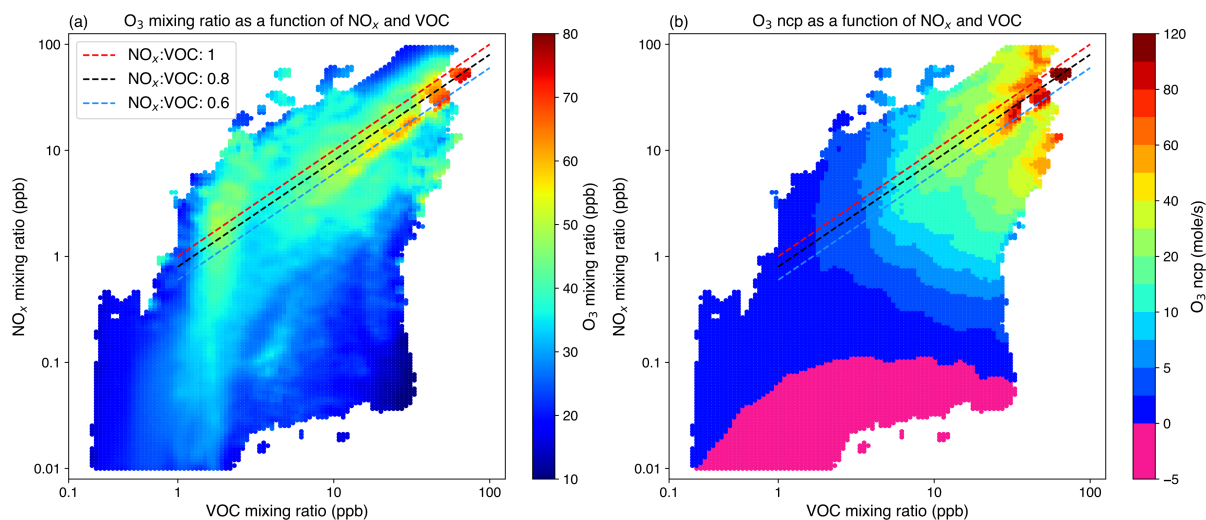
260 Non-linearity in chemical O<sub>3</sub> formation can result in differences in the effectiveness of emission control strategies regionally, and may aggravate O<sub>3</sub> pollution issues. We therefore investigate O<sub>3</sub> sensitivity for the present day and the future, as it is important to understand how O<sub>3</sub> mixing ratio concentrations will respond to changing emissions. Ratios of NO<sub>x</sub> and VOC mixing ratio concentrations provide a useful indicator of regional O<sub>3</sub> sensitivity regimes. Here we quantify the critical NO<sub>x</sub>/VOC ratio that distinguishes VOC-limited and NO<sub>x</sub>-limited regimes by examining monthly mean surface O<sub>3</sub> mixing ratio concentrations

265 and net chemical production rates as a function of monthly mean  $\text{NO}_x$  and VOC mixing ratio concentrations, see Fig. 6. Monthly mean data  $\text{O}_3$  mixing ratios and net production rates in the lowest model layer from all months and all scenarios are used to plot the isopleths figure. For  $\text{NO}_x$  we use the sum of  $\text{NO}$  and  $\text{NO}_2$  mixing ratios, and for total VOC we use the sum of the mixing ratios of primary emitted VOC species.  $\text{CO}$  and  $\text{CH}_4$  are not included due to their relatively low reactivity. We classify  $\text{NO}_x$  and VOC mixing ratios in each model grid cell into 150 bins on a logarithmic scale ranging from 0.01 ppb to 100

270 ppb, and calculate mean  $\text{O}_3$  mixing ratios and mean  $\text{O}_3$  net chemical production rates in each  $\text{NO}_x$ -VOC bin. Approximate thresholds of monthly mean  $\text{NO}_x/\text{VOC}$  ratios for  $\text{O}_3$  sensitivity are shown in Fig. 6, ranging from 0.6 to 1 with a central value of 0.8. We hence apply a threshold value of 0.8 to distinguish  $\text{O}_3$  sensitivity regimes hereafter. Areas above the threshold represent VOC-limited regimes in which increased  $\text{NO}_x$  emissions reduce  $\text{O}_3$  mixing ratio concentrations and  $\text{O}_3$  production rates, and areas below the threshold represent  $\text{NO}_x$ -limited regimes. We find the highest  $\text{O}_3$  mixing ratio concentrations and

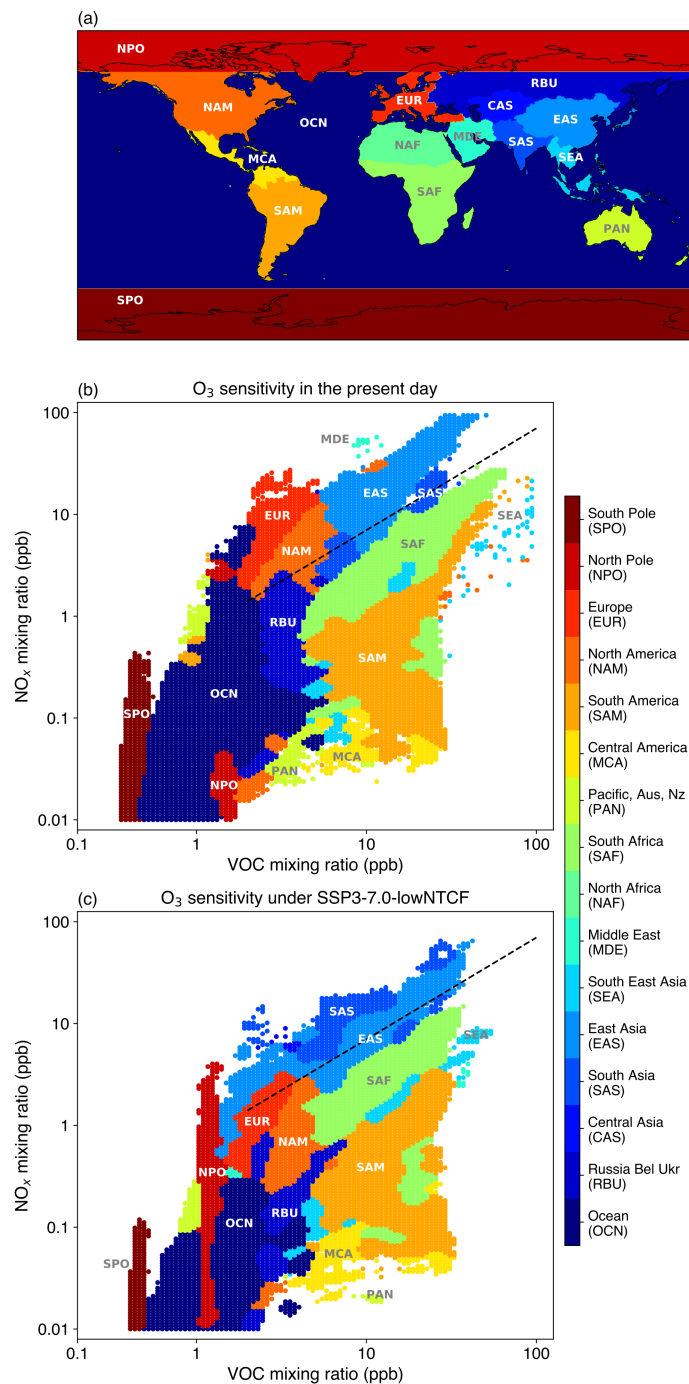
275  $\text{O}_3$  production close to the threshold line, demonstrating that the threshold values are consistent and that the approach is robust.

We further investigate  $\text{O}_3$  sensitivity for different regions. The regions considered here are those defined for the Task Force on Hemispheric Transport of Air Pollutants Phase 2 (TF HTAP2; Janssens-Maenhout et al., 2015), see Fig. 7a; the regions dominating each part of the  $\text{NO}_x$ -VOC mixing ratio concentration space are shown in Fig. 7b, c. We determine the dominant region contributing to each bin in  $\text{NO}_x$ -VOC space based on the region contributing the greatest number of model grid cells to that bin. ~~We identify the dominant region based on the number of model grid cells with the corresponding  $\text{NO}_x$  and VOC mixing ratios.~~ This approach reveals differences in regional  $\text{O}_3$  sensitivity. We also show the shift in  $\text{O}_3$  sensitivity in different regions between the present day (Fig. 7b) and the future (SSP3-7.0-lowNCTCF; Fig. 7c) to demonstrate the impacts of decreased NCTCF emissions on the evolution of  $\text{O}_3$  sensitivity on a regional basis.



**Figure 6.** Surface  $\text{O}_3$  mixing ratio concentrations (a) and  $\text{O}_3$  net chemical production (ncp) rates (b) as a function of monthly mean  $\text{NO}_x$  and VOC mixing ratio concentrations. Monthly mean data for all months and all scenarios are used. The straight lines show the approximate thresholds of  $\text{NO}_x/\text{VOC}$  to distinguish VOC-limited (above the line) and  $\text{NO}_x$ -limited (below the line) regimes.



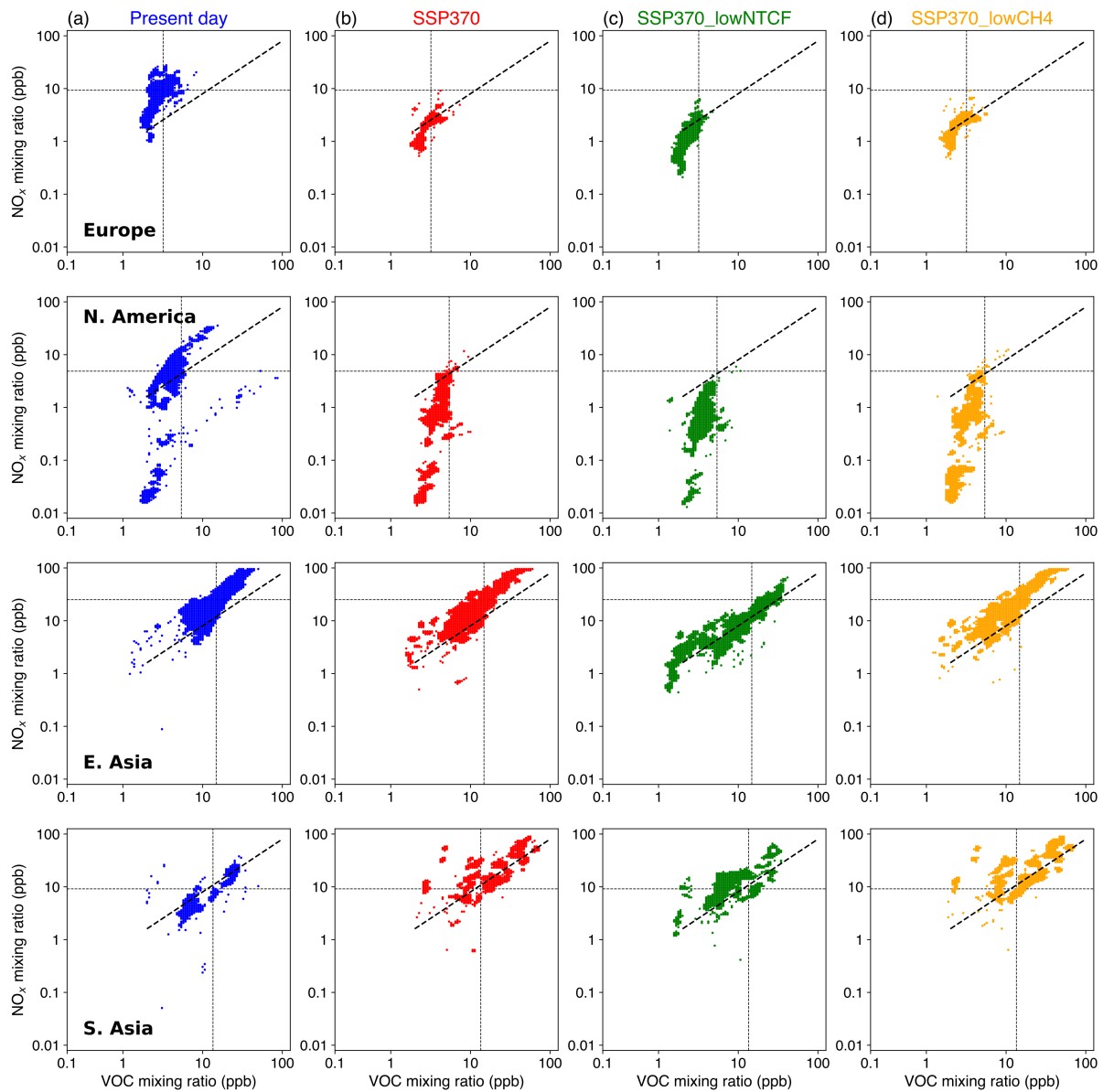


**Figure 7.** Geographical regions defined in the Task Force on Hemispheric Transport of Air Pollutants Phase 2 (TF HTAP2) (a). Regions dominating each part of NO<sub>x</sub>-VOC mixing ratio concentration space in the present day (b) and under SSP3-7.0-lowNTCF (c). The straight line shows the NO<sub>x</sub>/VOC threshold of 0.8. Monthly mean data for all months under each pathway are used to plot (b) and (c).

Figure 7b clearly shows that low-NO<sub>x</sub> and low-VOC environments are most common in oceanic and polar regions where surface O<sub>3</sub> levels are typically low and where O<sub>3</sub> production is NO<sub>x</sub>-limited. In contrast, Europe, North America and East Asia dominate the high-NO<sub>x</sub> and high-VOC environments in the present day where O<sub>3</sub> levels are high and O<sub>3</sub> production is VOC-limited. Europe and North America have similar VOC mixing ratio concentrations but NO<sub>x</sub> mixing ratio concentrations are generally higher in Europe than in North America, which results in Europe lying further above the NO<sub>x</sub>/VOC threshold. This demonstrates that stricter controls on NO<sub>x</sub> emissions are required for Europe to shift from VOC-limited to NO<sub>x</sub>-limited regimes than for North America. East Asia dominates VOC-limited regimes due to much higher NO<sub>x</sub> and VOC mixing ratios concentrations than other regions. Parts of South Asia and Middle East are also VOC-limited. Major biogenic emission source regions such as South America, South Africa and South East Asia have high VOC mixing ratio concentrations but moderate levels of NO<sub>x</sub>, and the chemical environment is therefore NO<sub>x</sub>-limited.

The impacts of reductions in NTCF emissions are shown in Fig. 7c. We find that Europe and North America are no longer the most dominant VOC-limited regimes due to decreased NO<sub>x</sub> mixing ratio concentrations. East Asia is still a dominant VOC-limited region under SSP3-7.0-lowNTCF, but reduced NO<sub>x</sub> emissions shift parts of East Asia into NO<sub>x</sub>-limitation. South Asia becomes the main VOC-limited region with relatively high NO<sub>x</sub> mixing ratio concentrations. South America, South Africa and South East Asia are still NO<sub>x</sub>-limited because there are no large NO<sub>x</sub> increases in these regions.

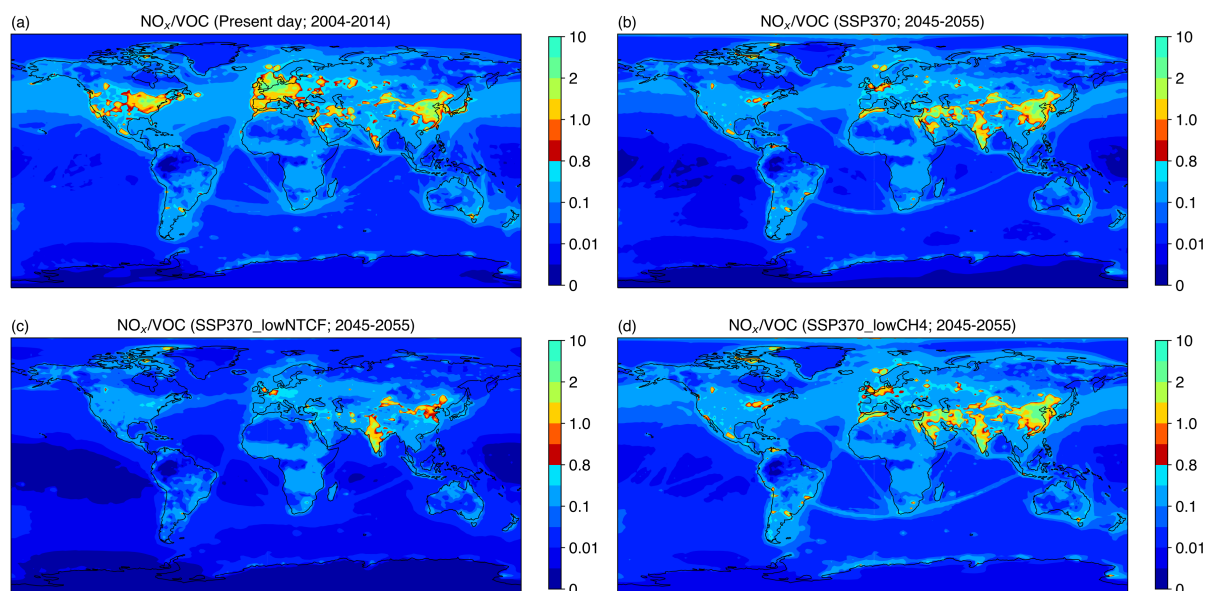
O<sub>3</sub> sensitivity in major present-day VOC-limited regions under different scenarios are shown in Fig. 8. Reductions in NO<sub>x</sub> emissions are important and effective in transforming VOC-limitation to NO<sub>x</sub>-limitation, reflected in large NO<sub>x</sub>-limited areas in Europe and North America under all scenarios. In contrast, most parts of East Asia are VOC-limited under SSP3-7.0 and SSP3-7.0-lowCH<sub>4</sub> due to increased NO<sub>x</sub> emissions. Since changes in VOC mixing ratio concentrations are relatively small for all scenarios, they have little substantial influence on O<sub>3</sub> sensitivity. Under future climate, increased biogenic VOC emissions in Europe and North America partly offset decreased anthropogenic and biomass VOC emissions, and are hence beneficial to maintain a NO<sub>x</sub>-limited environment. We note that South Asia is the only region that is substantially VOC-limited in the future due to increased NO<sub>x</sub> emissions. Reductions in CH<sub>4</sub> mixing ratio concentrations have relatively little impact on O<sub>3</sub> sensitivity, although they greatly reduce surface O<sub>3</sub> mixing ratio concentrations (Sect. 4.3).



**Figure 8.** O<sub>3</sub> sensitivity across Europe, North America, East Asia and South Asia in the present day (a) and under SSP3-7.0 (b), SSP3-7.0-lowNTCF (c) and SSP3-7.0-lowCH4 (d). The NO<sub>x</sub>/VOC threshold of 0.8 is shown. Horizontal and vertical lines indicate regional mean NO<sub>x</sub> and VOC mixing ratio concentrations in the present day.

## 6 Spatial distributions of O<sub>3</sub> sensitivity

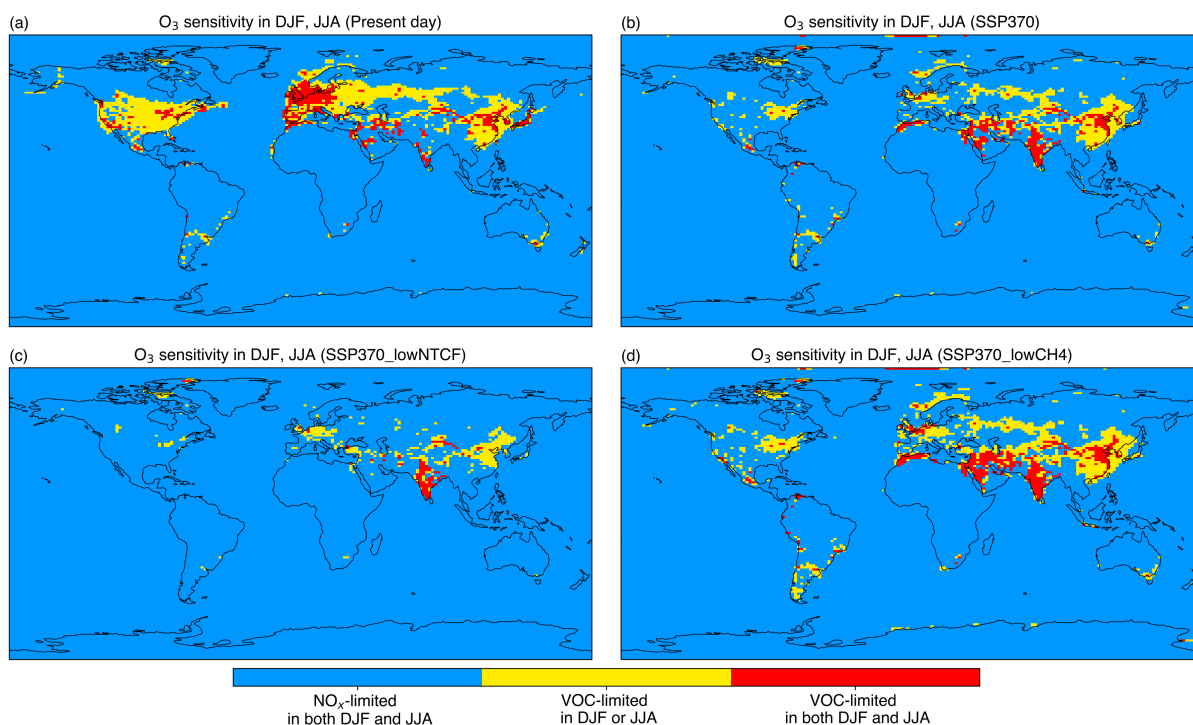
Global spatial distributions of annual O<sub>3</sub> sensitivity in the present day and the future are shown in Fig. 9. VOC-limited regimes are represented by high NO<sub>x</sub>/VOC ratios, reflecting relatively high NO<sub>x</sub> or low VOC mixing ratio concentrations. In the present day, high NO<sub>x</sub> emissions contribute to VOC-limitation in large areas of North America, Western and Central Europe and East Asia. While North America and Europe have lower NO<sub>x</sub> mixing ratio concentrations than East Asia, lower VOC mixing ratio concentrations still lead to VOC-limitation. Only south-west parts of India are VOC-limited. In the future, O<sub>3</sub> production in more areas of North America and Europe becomes NO<sub>x</sub>-limited. However, VOC limited regimes in East Asia, particularly China, are persistent due to projected increases in NO<sub>x</sub> emissions until 2055 under SSP3-7.0 and SSP3-7.0-lowCH<sub>4</sub>. We find that reductions in CH<sub>4</sub> mixing ratio concentrations have relatively little influence on O<sub>3</sub> sensitivity over continental regions (Fig. 9b vs 9d). We note that SSP3-7.0-lowNTCF shows the smallest VOC-limited areas across the globe. East Asia is still partly VOC-limited under this scenario, particularly in northern China (Fig. 9c), which indicates that further reductions in NO<sub>x</sub> emissions are required in addition to those expected in this region in the future (Fig. 3b).



**Figure 9.** Spatial distributions of annual mean NO<sub>x</sub>/VOC in the present day (2004–2014; **a**) and under SSP3-7.0 (**b**), SSP3-7.0-lowNTCF (**c**) and SSP3-7.0-lowCH<sub>4</sub> (**d**) (2045–2055). A NO<sub>x</sub>/VOC threshold ratio of 0.8 is used here to distinguish O<sub>3</sub> sensitivity regimes.

320 We contrast regional O<sub>3</sub> sensitivities for winter and summer seasons in Fig 10, as we note that O<sub>3</sub> responses in different seasons can be substantially different in high-emission regions. This suggests that static emission control strategies throughout the year may not be the best way to lower annual-mean O<sub>3</sub> pollution, and adjustments may be needed according to seasonal O<sub>3</sub> sensitivities. More extensive VOC-limited areas are found in winter than in summer under both present day and future conditions, but these account for less than 7 % of the total area of the world in the present day (Table 5). Over 50 % of North

325 America, Europe and East Asia are VOC-limited in winter in the present day. This explains why North America and Europe show  $O_3$  increases in winter (Fig. 4) despite reduced  $NO_x$  emissions. However, there are fewer VOC-limited regions across the globe under all future pathways (Fig. 10). About 1 % of North America is VOC-limited in summer, and less than 7 % of Europe. In contrast, over 48 % (winter) and 39 % (summer) of South Asia is VOC-limited in future. Slightly more areas are VOC-limited under SSP3-7.0-lowCH4 than under SSP3-7.0 (Fig 10c, d). Overall, reductions in  $NO_x$  emissions are important to reduce  $O_3$  production in high-emission regions and shift VOC-limited areas to  $NO_x$ -limitation but this may lead to higher  $O_3$  mixing ratio concentrations in winter without further emission controls on VOC and  $CH_4$ .



**Figure 10.** Seasonal differences in  $O_3$  sensitivity regimes for the present day (2004–2014; **a**) and under SSP3-7.0 (**b**), SSP3-7.0-lowNTCF (**c**), SSP3-7.0-lowCH4 (**d**) (2045–2055). A  $NO_x/VOC$  threshold ratio of 0.8 is used here to distinguish  $O_3$  sensitivity regimes.

**Table 5.** The percentage of VOC-limited areas (%) in different regions under different pathways in winter (DJF) and in summer (JJA).

	Present day		SSP3-7.0		SSP3-7.0-lowNTCF		SSP3-7.0-lowCH4	
	DJF	JJA	DJF	JJA	DJF	JJA	DJF	JJA
N. America	50.4	6.4	12.5	0.4	2.4	0	16.9	1.0
Europe	79.8	37.4	27.0	2.6	27.5	0.9	41.6	6.1
S. Asia	15.1	15.3	54.8	49.6	48.1	39.5	60.4	56.7
E. Asia	60.3	18.3	63.3	15.8	37.7	3.5	68.7	19.3
Globe	6.9	2.7	4.8	2.0	2.0	0.6	5.9	3.0

## 7 Conclusions

We use a global chemistry-climate model, UKESM1, to assess the impacts of changing near-term climate forcer (NTCF) emissions and CH<sub>4</sub> [mixing ratio concentrations](#) in the context of climate change on tropospheric O<sub>3</sub> in the present day (2004–2014) and the near future (2045–2055). CMIP6 future scenarios [including representing ‘regional rivalry’ development pathways](#) (SSP3-7.0, SSP3-7.0-lowNTCF and SSP3-7.0-lowCH4) are used from the AerChemMIP project. We have examined O<sub>3</sub> changes from the present day to the future and investigated regional O<sub>3</sub> sensitivities to explain contrasting O<sub>3</sub> changes in different seasons.

An extended chemistry scheme incorporating more reactive VOC species is used to permit representation of more active photochemical environments for O<sub>3</sub> production. This shows higher surface O<sub>3</sub> [mixing ratio concentrations](#) in high-emission regions and a 5 % higher tropospheric O<sub>3</sub> burden. While simulated surface O<sub>3</sub> [mixing ratio concentrations](#) are biased low in winter and high in summer, these systematic model biases are similar to those using the original chemistry scheme as well as a more comprehensive chemistry scheme. This indicates that other factors in the model are likely to be responsible for the biases but the extended chemistry scheme permits representation of a more appropriate chemical environment for O<sub>3</sub> production in high-emission areas.

From the present day to the future, the tropospheric O<sub>3</sub> burden increases by 4 % under SSP3-7.0 and decreases by 7 % and 5 % under SSP3-7.0-lowNTCF and SSP3-7.0-lowCH4. The tropospheric O<sub>3</sub> chemical lifetime remains similar (21–22 days) under all scenarios, and this is similar to previous estimates. Seasonal global mean surface O<sub>3</sub> [mixing ratio concentrations](#) increase by 0.7–0.9 ppb under SSP3-7.0, and decrease by 3.3–5.2 ppb under SSP3-7.0-lowNTCF and by 2.7–3.5 ppb under SSP3-7.0-lowCH4. We find that reductions in NTCF emissions are effective in reducing surface O<sub>3</sub> [mixing ratio concentrations](#), and reductions in CH<sub>4</sub> [mixing ratio concentrations](#) are also important. We also find that both the magnitude and direction of seasonal, daytime and nighttime O<sub>3</sub> changes relative to the present day can vary greatly across different regions especially South and East Asia.

O<sub>3</sub> sensitivity is quantified using monthly mean NO<sub>x</sub>/VOC **concentration** ratios to give a broad assessment of regional  
355 O<sub>3</sub> sensitivity. The estimated monthly mean NO<sub>x</sub>/VOC thresholds range from 0.6 to 1.0, and 0.8 is used to distinguish O<sub>3</sub>  
sensitivity regimes. Most VOC limited regimes occur in high-emission regions across the northern hemisphere, such as North  
America, Europe, the Middle East, South Asia and East Asia. More areas in North America and Europe become NO<sub>x</sub>-limited  
under all future pathways due to the projected decrease in NO<sub>x</sub> emissions. There are more VOC-limited areas in East Asia  
360 areas in East Asia under SSP3-7.0 and SSP3-7.0-lowCH<sub>4</sub> due to the projected increase in NO<sub>x</sub> emissions, although there are fewer VOC-limited  
areas in East Asia under SSP3-7.0-lowNTCF. South Asia becomes the dominant region for VOC-limited O<sub>3</sub> production in the  
future. Projections of regional O<sub>3</sub> sensitivity demonstrate that reductions in NO<sub>x</sub> emissions are the most important factor to  
shift VOC-limited regimes to NO<sub>x</sub>-limitation.

We highlight that O<sub>3</sub> sensitivity varies by season. There are more VOC-limited regimes in winter (7 %) than in summer (3  
%) across the globe. Reductions in NO<sub>x</sub> emissions increase surface O<sub>3</sub> **mixing ratio concentrations** in high-emission areas par-  
365 ticularly in winter, and reductions in VOC emissions should be targeted. In the future, reductions in NO<sub>x</sub> and VOC emissions  
should both be effective in mitigating O<sub>3</sub> pollution in most areas of North America and Europe in summer because there are  
only 1 % and 7 % VOC-limited areas in these two regions. However, further reductions in NO<sub>x</sub> emissions are needed for parts  
of East Asia and South Asia to convert most VOC-limited areas to NO<sub>x</sub>-limited. While anthropogenic and biomass emissions  
may be controlled in the future, more biogenic emissions under a warmer climate would hinder the impacts of VOCs on O<sub>3</sub>  
370 mitigation. Reductions in CH<sub>4</sub> **mixing ratio concentrations** are also important to reduce surface O<sub>3</sub> pollution. NO<sub>x</sub> decreases  
are important to reduce surface O<sub>3</sub> **mixing ratio concentrations** from a **global globe** perspective but will lead to increased O<sub>3</sub>  
**mixing ratio concentrations** in some regions, and hence emission controls on VOC and CH<sub>4</sub> are necessary to mitigate regional  
O<sub>3</sub> pollution during the transition from VOC- to NO<sub>x</sub>-limitation.

*Data availability.* The data generated in this study are available upon request.

375 *Author contributions.* ZL, RD, OW, FO'C, ST designed the study. ZL set up the model, conducted model simulations and performed the  
analysis. ZL, RD and OW prepared the paper. All co-authors contributed to reviewing and editing the paper.

*Competing interests.* The authors declare that they have no conflict of interest.

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## References

- Agathokleous, E., Feng, Z. Z., Oksanen, E., Sicard, P., Wang, Q., Saitanis, C. J., Araminiene, V., Blande, J. D., Hayes, F., Calatayud, V., Domingos, M., Veresoglou, S. D., Penuelas, J., Wardle, D. A., De Marco, A., Li, Z. Z., Harmens, H., Yuan, X. Y., Vitale, M., and Paoletti, E.: Ozone affects plant, insect, and soil microbial communities: A threat to terrestrial ecosystems and biodiversity, *Science Advances*, 6, <https://doi.org/10.1126/sciadv.abc1176>, 2020.
- Allen, R. J., Horowitz, L. W., Naik, V., Oshima, N., O'Connor, F. M., Turnock, S., Shim, S., Le Sager, P., van Noije, T., Tsigaridis, K., Bauer, S. E., Sentman, L. T., John, J. G., Broderick, C., Deushi, M., Folberth, G. A., Fujimori, S., and Collins, W. J.: Significant climate benefits from near-term climate forcer mitigation in spite of aerosol reductions, *Environmental Research Letters*, 16, <https://doi.org/10.1088/1748-9326/abe06b>, 2021.
- Archer-Nicholls, S., Abraham, N. L., Shin, Y. M., Weber, J., Russo, M. R., Lowe, D., Utembe, S., O'Connor, F. M., Kerridge, B., Latter, B., et al.: The Common Representative Intermediates Mechanism version 2 in the United Kingdom Chemistry and Aerosols Model, <https://doi.org/10.1029/2020MS002420>, 2020.
- Archibald, A. T., O'Connor, F. M., Abraham, N. L., Archer-Nicholls, S., Chipperfield, M. P., Dalvi, M., Folberth, G. A., Dennison, F., Dhomse, S. S., Griffiths, P. T., Hardacre, C., Hewitt, A. J., Hill, R. S., Johnson, C. E., Keeble, J., Kohler, M. O., Morgenstern, O., Mulcahy, J. P., Ordonez, C., Pope, R. J., Rumbold, S. T., Russo, M. R., Savage, N. H., Sellar, A., Stringer, M., Turnock, S. T., Wild, O., and Zeng, G.: Description and evaluation of the UKCA stratosphere-troposphere chemistry scheme (StratTrop vn 1.0) implemented in UKESM1, *Geoscientific Model Development*, 13, 1223–1266, <https://doi.org/10.5194/gmd-13-1223-2020>, 2020a.
- Archibald, A. T., Neu, J. L., Elshorbany, Y. F., Cooper, O. R., Young, P. J., Akiyoshi, H., Cox, R. A., Coyle, M., Derwent, R. G., Deushi, M., Finco, A., Frost, G. J., Galbally, I. E., Gerosa, G., Granier, C., Griffiths, P. T., Hossaini, R., Hu, L., Jockel, P., Josse, B., Lin, M. Y., Mertens, M., Morgenstern, O., Naja, M., Naik, V., Oltmans, S., Plummer, D. A., Revell, L. E., Saiz-Lopez, A., Saxena, P., Shin, Y. M., Shahid, I., Shallcross, D., Tilmes, S., Trickl, T., Wallington, T. J., Wang, T., Worden, H. M., and Zeng, G.: Tropospheric Ozone Assessment Report: A critical review of changes in the tropospheric ozone burden and budget from 1850 to 2100, *Elementa-Science of the Anthropocene*, 8, <https://doi.org/10.1525/elementa.2020.034>, 2020b.
- Bieser, J., Aulinger, A., Matthias, V., Quante, M., and Van Der Gon, H. D.: Vertical emission profiles for Europe based on plume rise calculations, 159, 2935–2946, <https://doi.org/10.1016/j.envpol.2011.04.030>, 2011.
- Colette, A., Aas, W., Banin, L., Braban, C. F., Ferm, M., Gonzalez Ortiz, A., Ilyin, I., Mar, K., Pandolfi, M., and Putaud, J.-P.: Air pollution trends in the EMEP region between 1990 and 2012, <http://nora.nerc.ac.uk/id/eprint/513779>, 2016.
- Collins, W. J., Lamarque, J. F., Schulz, M., Boucher, O., Eyring, V., Hegglin, M. I., Maycock, A., Myhre, G., Prather, M., Shindell, D., and Smith, S. J.: AerChemMIP: quantifying the effects of chemistry and aerosols in CMIP6, *Geoscientific Model Development*, 10, 585–607, <https://doi.org/10.5194/gmd-10-585-2017>, 2017.
- Doherty, R. M., Wild, O., Shindell, D. T., Zeng, G., MacKenzie, I. A., Collins, W. J., Fiore, A. M., Stevenson, D. S., Dentener, F. J., Schultz, M. G., Hess, P., Derwent, R. G., and Keating, T. J.: Impacts of climate change on surface ozone and intercontinental ozone pollution: A multi-model study, *Journal of Geophysical Research-Atmospheres*, 118, 3744–3763, <https://doi.org/10.1002/jgrd.50266>, 2013.
- Doherty, R. M., Heal, M. R., and O'Connor, F. M.: Climate change impacts on human health over Europe through its effect on air quality, *Environmental Health*, 16, 33–44, 2017.
- Dunker, A. M., Yarwood, G., Ortmann, J. P., and Wilson, G. M.: Comparison of source apportionment and source sensitivity of ozone in a three-dimensional air quality model, *Environmental science & technology*, 36, 2953–2964, 2002.

- 420 Eyring, V., Bony, S., Meehl, G. A., Senior, C. A., Stevens, B., Stouffer, R. J., and Taylor, K. E.: Overview of the Coupled Model Intercomparison Project Phase 6 (CMIP6) experimental design and organization, *Geosci. Model Dev.*, 9, 1937–1958, <https://doi.org/10.5194/gmd-9-1937-2016>, 2016.
- Fenech, S., Doherty, R. M., Heaviside, C., Vardoulakis, S., Macintyre, H. L., and O'Connor, F. M.: The influence of model spatial resolution on simulated ozone and fine particulate matter for Europe: implications for health impact assessments, *Atmospheric Chemistry and Physics*, 18, 5765–5784, <https://doi.org/10.5194/acp-18-5765-2018>, 2018.
- 425 Feng, L. Y., Smith, S. J., Braun, C., Crippa, M., Gidden, M. J., Hoesly, R., Klimont, Z., van Marle, M., van den Berg, M., and van der Werf, G. R.: The generation of gridded emissions data for CMIP6, *Geoscientific Model Development*, 13, 461–482, <https://doi.org/10.5194/gmd-13-461-2020>, 2020.
- Fiore, A. M., West, J. J., Horowitz, L. W., Naik, V., and Schwarzkopf, M. D.: Characterizing the tropospheric ozone response to methane emission controls and the benefits to climate and air quality, *Journal of Geophysical Research-Atmospheres*, 113, <https://doi.org/10.1029/2007jd009162>, 2008.
- 430 Fiore, A. M., Naik, V., Spracklen, D. V., Steiner, A., Unger, N., Prather, M., Bergmann, D., Cameron-Smith, P. J., Cionni, I., Collins, W. J., Dalsøren, S., Eyring, V., Folberth, G. A., Ginoux, P., Horowitz, L. W., Josse, B., Lamarque, J.-F., MacKenzie, I. A., Nagashima, T., O'Connor, F. M., Righi, M., Rumbold, S. T., Shindell, D. T., Skeie, R. B., Sudo, K., Szopa, S., Takemura, T., and Zeng, G.: Global air quality and climate, *Chemical Society Reviews*, 41, 6663–6683, <https://doi.org/10.1039/C2CS35095E>, 2012.
- 435 Fujimori, S., Hasegawa, T., Masui, T., Takahashi, K., Herran, D. S., Dai, H. C., Hijioka, Y., and Kainuma, M.: SSP3: AIM implementation of Shared Socioeconomic Pathways, *Global Environmental Change-Human and Policy Dimensions*, 42, 268–283, <https://doi.org/10.1016/j.gloenvcha.2016.06.009>, 2017.
- Gao, M., Gao, J., Zhu, B., Kumar, R., Lu, X., Song, S., Zhang, Y., Jia, B., Wang, P., Beig, G., Hu, J., Ying, Q., Zhang, H., Sherman, P., and McElroy, M. B.: Ozone pollution over China and India: seasonality and sources, *Atmos. Chem. Phys.*, 20, 4399–4414, <https://doi.org/10.5194/acp-20-4399-2020>, 2020.
- 440 Griffiths, P. T., Murray, L. T., Zeng, G., Shin, Y. M., Abraham, N. L., Archibald, A. T., Deushi, M., Emmons, L. K., Galbally, I. E., Hassler, B., Horowitz, L. W., Keeble, J., Liu, J., Moeini, O., Naik, V., O'Connor, F. M., Oshima, N., Tarasick, D., Tilmes, S., Turnock, S. T., Wild, O., Young, P. J., and Zanis, P.: Tropospheric ozone in CMIP6 simulations, *Atmos. Chem. Phys.*, 21, 4187–4218, <https://doi.org/10.5194/acp-21-4187-2021>, 2021.
- 445 Hakim, Z. Q., Archer-Nicholls, S., Beig, G., Folberth, G. A., Sudo, K., Abraham, N. L., Ghude, S., Henze, D. K., and Archibald, A. T.: Evaluation of tropospheric ozone and ozone precursors in simulations from the HTAPII and CCMI model intercomparisons—a focus on the Indian subcontinent, *Atmospheric Chemistry and Physics*, 19, 6437–6458, 2019.
- Hoesly, R. M., Smith, S. J., Feng, L. Y., Klimont, Z., Janssens-Maenhout, G., Pitkanen, T., Seibert, J. J., Vu, L., Andres, R. J., Bolt, R. M., Bond, T. C., Dawidowski, L., Kholod, N., Kurokawa, J., Li, M., Liu, L., Lu, Z. F., Moura, M. C. P., O'Rourke, P. R., and Zhang, Q.: Historical (1750–2014) anthropogenic emissions of reactive gases and aerosols from the Community Emissions Data System (CEDS), *Geoscientific Model Development*, 11, 369–408, <https://doi.org/10.5194/gmd-11-369-2018>, 2018.
- 450 Huang, G., Brook, R., Crippa, M., Janssens-Maenhout, G., Schieberle, C., Dore, C., Guizzardi, D., Muntean, M., Schaaf, E., and Friedrich, R.: Speciation of anthropogenic emissions of non-methane volatile organic compounds: a global gridded data set for 1970–2012, *Atmos. Chem. Phys.*, 17, 7683–7701, <https://doi.org/10.5194/acp-17-7683-2017>, 2017.
- 455 Janssens-Maenhout, G., Crippa, M., Guizzardi, D., Dentener, F., Muntean, M., Pouliot, G., Keating, T., Zhang, Q., Kurokawa, J., Wankmüller, R., Denier van der Gon, H., Kuenen, J. J. P., Klimont, Z., Frost, G., Darras, S., Koffi, B., and Li, M.: HTAP\_v2.2: a mosaic of regional

- and global emission grid maps for 2008 and 2010 to study hemispheric transport of air pollution, *Atmos. Chem. Phys.*, 15, 11 411–11 432, <https://doi.org/10.5194/acp-15-11411-2015>, 2015.
- 460 Jin, X. M. and Holloway, T.: Spatial and temporal variability of ozone sensitivity over China observed from the Ozone Monitoring Instrument, *Journal of Geophysical Research-Atmospheres*, 120, 7229–7246, <https://doi.org/10.1002/2015jd023250>, 2015.
- Kleinman, L. I.: Low and high NO<sub>x</sub> tropospheric photochemistry, *Journal of Geophysical Research Atmospheres*, 99, 16 831–16 838, 1994.
- Lamarque, J. F., Bond, T. C., Eyring, V., Granier, C., Heil, A., Klimont, Z., Lee, D., Liousse, C., Mieville, A., Owen, B., Schultz, M. G., Shindell, D., Smith, S. J., Stehfest, E., Van Aardenne, J., Cooper, O. R., Kainuma, M., Mahowald, N., McConnell, J. R., Naik, V., Riahi, K., and van Vuuren, D. P.: Historical (1850–2000) gridded anthropogenic and biomass burning emissions of reactive gases and aerosols: methodology and application, *Atmos. Chem. Phys.*, 10, 7017–7039, <https://doi.org/10.5194/acp-10-7017-2010>, 2010.
- 465 Lefohn, A. S., Malley, C. S., Smith, L., Wells, B., Hazucha, M., Simon, H., Naik, V., Mills, G., Schultz, M. G., Paoletti, E., De Marco, A., Xu, X. B., Zhang, L., Wang, T., Neufeld, H. S., Musselman, R. C., Tarasick, D., Brauer, M., Feng, Z. Z., Tang, H. Y., Kobayashi, K., Sicard, P., Solberg, S., and Gerosa, G.: Tropospheric ozone assessment report: Global ozone metrics for climate change, human health, and crop/ecosystem research, *Elementa-Science of the Anthropocene*, 6, <https://doi.org/10.1525/elementa.279>, 2018.
- 470 Lelieveld, J. and Dentener, F. J.: What controls tropospheric ozone?, 105, 3531–3551, <https://doi.org/10.1029/1999JD901011>, 2000.
- Lin, M., Horowitz, L. W., Xie, Y., Paulot, F., Malyshev, S., Shevliakova, E., Finco, A., Gerosa, G., Kubistin, D., and Pilegaard, K.: Vegetation feedbacks during drought exacerbate ozone air pollution extremes in Europe, *Nature Climate Change*, 10, 444–451, 2020.
- Liu, Z., Doherty, R. M., Wild, O., Holloway, M., and O’Connor, F. M.: Contrasting chemical environments in summertime for atmospheric ozone across major Chinese industrial regions: the effectiveness of emission control strategies, *Atmos. Chem. Phys.*, 21, 10 689–10 706, <https://doi.org/10.5194/acp-21-10689-2021>, 2021.
- 475 Lu, X., Zhang, L., Liu, X., Gao, M., Zhao, Y., and Shao, J.: Lower tropospheric ozone over India and its linkage to the South Asian monsoon, *Atmos. Chem. Phys.*, 18, 3101–3118, <https://doi.org/10.5194/acp-18-3101-2018>, 2018.
- Lu, X., Zhang, L., Wang, X., Gao, M., Li, K., Zhang, Y., Yue, X., and Zhang, Y.: Rapid increases in warm-season surface ozone and resulting health impact in China since 2013, *Environmental Science & Technology Letters*, 7, 240–247, 2020.
- 480 Ma, M., Gao, Y., Wang, Y., Zhang, S., Leung, L. R., Liu, C., Wang, S., Zhao, B., Chang, X., Su, H., et al.: Substantial ozone enhancement over the North China Plain from increased biogenic emissions due to heat waves and land cover in summer 2017, *Atmospheric Chemistry and Physics*, 19, 12 195–12 207, 2019.
- Mailler, S., Khvorostyanov, D., and Menut, L.: Impact of the vertical emission profiles on background gas-phase pollution simulated from the EMEP emissions over Europe, *Atmospheric Chemistry and Physics*, 13, 5987–5998, <https://doi.org/10.5194/acp-13-5987-2013>, 2013.
- 485 Mann, G. W., Carslaw, K. S., Spracklen, D. V., Ridley, D. A., Manktelow, P. T., Chipperfield, M. P., Pickering, S. J., and Johnson, C. E.: Description and evaluation of GLOMAP-mode: a modal global aerosol microphysics model for the UKCA composition-climate model, *Geosci. Model Dev.*, 3, 519–551, <https://doi.org/10.5194/gmd-3-519-2010>, 2010.
- Meinshausen, M., Vogel, E., Nauels, A., Lorbacher, K., Meinshausen, N., Etheridge, D. M., Fraser, P. J., Montzka, S. A., Rayner, P. J., Trudinger, C. M., Krummel, P. B., Beyerle, U., Canadell, J. G., Daniel, J. S., Enting, I. G., Law, R. M., Lunder, C. R., O’Doherty, S., Prinn, R. G., Reimann, S., Rubino, M., Velders, G. J. M., Vollmer, M. K., Wang, R. H. J., and Weiss, R.: Historical greenhouse gas concentrations for climate modelling (CMIP6), *Geosci. Model Dev.*, 10, 2057–2116, <https://doi.org/10.5194/gmd-10-2057-2017>, 2017.
- 490 Meinshausen, M., Nicholls, Z. R. J., Lewis, J., Gidden, M. J., Vogel, E., Freund, M., Beyerle, U., Gessner, C., Nauels, A., Bauer, N., Canadell, J. G., Daniel, J. S., John, A., Krummel, P. B., Luderer, G., Meinshausen, N., Montzka, S. A., Rayner, P. J., Reimann, S., Smith,

- 495 S. J., van den Berg, M., Velders, G. J. M., Vollmer, M. K., and Wang, R. H. J.: The shared socio-economic pathway (SSP) greenhouse gas concentrations and their extensions to 2500, *Geosci. Model Dev.*, 13, 3571–3605, <https://doi.org/10.5194/gmd-13-3571-2020>, 2020.
- Morgenstern, O., Braesicke, P., O'Connor, F. M., Bushell, A. C., Johnson, C. E., Osprey, S. M., and Pyle, J. A.: Evaluation of the new UKCA climate-composition model – Part 1: The stratosphere, *Geosci. Model Dev.*, 2, 43–57, <https://doi.org/10.5194/gmd-2-43-2009>, 2009.
- Mulcahy, J. P., Johnson, C., Jones, C. G., Povey, A. C., Scott, C. E., Sellar, A., Turnock, S. T., Woodhouse, M. T., Abraham, N. L., Andrews, M. B., Bellouin, N., Browse, J., Carslaw, K. S., Dalvi, M., Folberth, G. A., Glover, M., Grosvenor, D. P., Hardacre, C., Hill, R., Johnson, B., Jones, A., Kipling, Z., Mann, G., Mollard, J., O'Connor, F. M., Palmieri, J., Reddington, C., Rumbold, S. T., Richardson, M., Schutgens, N. A. J., Stier, P., Stringer, M., Tang, Y. M., Walton, J., Woodward, S., and Yool, A.: Description and evaluation of aerosol in UKESM1 and HadGEM3-GC3.1 CMIP6 historical simulations, *Geoscientific Model Development*, 13, 6383–6423, <https://doi.org/10.5194/gmd-13-6383-2020>, 2020.
- 500 Naik, V., Szopa, S., Adhikary, B., Artaxo, P., Bernsten, T., Collins, W., Fuzzi, S., Gallardo, L., Kiendler-Scharr, A., Klimont, Z., Liao, H., Unger, N., Zanis, P., et al.: *Climate Change 2021: The Physical Science Basis. Contribution of Working Group I to the Sixth Assessment Report of the Intergovernmental Panel on Climate Change: Chapter 6: Short-Lived Climate Forcers*, IPCC, 2021.
- O'Connor, F. M., Johnson, C. E., Morgenstern, O., Abraham, N. L., Braesicke, P., Dalvi, M., Folberth, G. A., Sanderson, M. G., Telford, P. J., Voulgarakis, A., Young, P. J., Zeng, G., Collins, W. J., and Pyle, J. A.: Evaluation of the new UKCA climate-composition model - Part 2: The Troposphere, *Geoscientific Model Development*, 7, 41–91, <https://doi.org/10.5194/gmd-7-41-2014>, 2014.
- 510 O'Connor, F. M., Abraham, N. L., Dalvi, M., Folberth, G. A., Griffiths, P. T., Hardacre, C., Johnson, B., Kahana, R., Keeble, J., Kim, B., Morgenstern, O., Mulcahy, J. P., Richardson, M., Robertson, E., Seo, J., Shim, S., Teixeira, J. C., Turnock, S. T., Williams, J., Wiltshire, A. J., Woodward, S., and Zeng, G.: Assessment of pre-industrial to present-day anthropogenic climate forcing in UKESM1, *Atmospheric Chemistry and Physics*, 21, 1211–1243, <https://doi.org/10.5194/acp-21-1211-2021>, 2021.
- 515 O'Neill, B. C., Kriegler, E., Riahi, K., Ebi, K. L., Hallegatte, S., Carter, T. R., Mathur, R., and van Vuuren, D. P.: A new scenario framework for climate change research: the concept of shared socioeconomic pathways, *Climatic Change*, 122, 387–400, <https://doi.org/10.1007/s10584-013-0905-2>, 2014.
- Pacifico, F., Harrison, S. P., Jones, C. D., Arneth, A., Sitch, S., Weedon, G. P., Barkley, M. P., Palmer, P. I., Serça, D., Potosnak, M., Fu, T. M., Goldstein, A., Bai, J., and Schurgers, G.: Evaluation of a photosynthesis-based biogenic isoprene emission scheme in JULES and simulation of isoprene emissions under present-day climate conditions, *Atmos. Chem. Phys.*, 11, 4371–4389, <https://doi.org/10.5194/acp-11-4371-2011>, 2011.
- 520 Prather, M., Flato, G., Friedlingstein, P., Jones, C., Lamarque, J.-F., Liao, H., Rasch, P., et al.: *Climate change 2013: the physical science basis: Working Group I contribution to the Fifth assessment report of the Intergovernmental Panel on Climate Change*, Cambridge university press, 2014.
- 525 Rao, S., Klimont, Z., Smith, S. J., Van Dingenen, R., Dentener, F., Bouwman, L., Riahi, K., Amann, M., Bodirsky, B. L., van Vuuren, D. P., Reis, L. A., Calvin, K., Drouet, L., Fricko, O., Fujimori, S., Gernaat, D., Havlik, P., Harmsen, M., Hasegawa, T., Heyes, C., Hilaire, J., Luderer, G., Masui, T., Stehfest, E., Strefler, J., van der Sluis, S., and Tavoni, M.: Future air pollution in the Shared Socio-economic Pathways, *Global Environmental Change-Human and Policy Dimensions*, 42, 346–358, <https://doi.org/10.1016/j.gloenvcha.2016.05.012>, 2017.
- 530 Romer, P. S., Duffey, K. C., Wooldridge, P. J., Edgerton, E., Baumann, K., Feiner, P. A., Miller, D. O., Brune, W. H., Koss, A. R., Gouw, J. A. d., et al.: Effects of temperature-dependent NO<sub>x</sub> emissions on continental ozone production, *Atmospheric Chemistry and Physics*, 18, 2601–2614, 2018.

- Schnell, J. L. and Prather, M. J.: Co-occurrence of extremes in surface ozone, particulate matter, and temperature over eastern North America, *Proceedings of the National Academy of Sciences*, 114, 2854–2859, 2017.
- 535 Schultz, M., Schröder, S., Lyapina, O., Cooper, O., Galbally, I., Petropavlovskikh, I., Von Schneidemesser, E., Tanimoto, H., Elshorbany, Y., Naja, M., et al.: Tropospheric Ozone Assessment Report, links to Global surface ozone datasets, PANGAEA, 2017a.
- Schultz, M. G., Schroder, S., Lyapina, O., Cooper, O. R., Galbally, I., Petropavlovskikh, I., von Schneidemesser, E., Tanimoto, H., Elshorbany, Y., Naja, M., Seguel, R. J., Dauert, U., Eckhardt, P., Feigenspan, S., Fiebig, M., Hjellbrekke, A. G., Hong, Y. D., Kjeld, P. C., Koide, H., Lear, G., Tarasick, D., Ueno, M., Wallasch, M., Baumgardner, D., Chuang, M. T., Gillett, R., Lee, M., Molloy, S., Moolla, R., Wang, T., Sharps, K., Adame, J. A., Ancellet, G., Apadula, F., Artaxo, P., Barlasina, M. E., Bogucka, M., Bonasoni, P., Chang, L., Colomb, A., Cuevas-Agullo, E., Cupeiro, M., Degorska, A., Ding, A. J., FrHlich, M., Frolova, M., Gadhavi, H., Gheusi, F., Gilge, S., Gonzalez, M. Y., Gros, V., Hamad, S. H., Helmig, D., Henriques, D., Hermansen, O., Holla, R., Hueber, J., Im, U., Jaffe, D. A., Komala, N., Kubistin, D., Lam, K. S., Laurila, T., Lee, H., Levy, I., Mazzoleni, C., Mazzoleni, L. R., McClure-Begley, A., Mohamad, M., Murovec, M., Navarro-Comas, M., Nicodim, F., Parrish, D., Read, K. A., Reid, N., Ries, N. R. L., Saxena, P., Schwab, J. J., Scorgie, Y., Senik, I., Simmonds, P., 540 Sinha, V., Skorokhod, A. I., Spain, G., Spangl, W., Spoor, R., Springston, S. R., Steer, K., Steinbacher, M., Suharguniyawan, E., Torre, P., Trickl, T., Lin, W. L., Weller, R., Xu, X. B., Xue, L. K., and Ma, Z. Q.: Tropospheric Ozone Assessment Report: Database and metrics data of global surface ozone observations, *Elementa-Science of the Anthropocene*, 5, <https://doi.org/10.1525/elementa.244>, 2017b.
- Sellar, A. A., Jones, C. G., Mulcahy, J. P., Tang, Y. M., Yool, A., Wiltshire, A., O'Connor, F. M., Stringer, M., Hill, R., Palmieri, J., Woodward, S., de Mora, L., Kuhlbrodt, T., Rumbold, S. T., Kelley, D. I., Ellis, R., Johnson, C. E., Walton, J., Abraham, N. L., Andrews, M. B., 550 Andrews, T., Archibald, A. T., Berthou, S., Burke, E., Blockley, E., Carslaw, K., Dalvi, M., Edwards, J., Folberth, G. A., Gedney, N., Griffiths, P. T., Harper, A. B., Hendry, M. A., Hewitt, A. J., Johnson, B., Jones, A., Jones, C. D., Keeble, J., Liddicoat, S., Morgenstern, O., Parker, R. J., Predoi, V., Robertson, E., Siahhaan, A., Smith, R. S., Swaminathan, R., Woodhouse, M. T., Zeng, G., and Zerroukat, M.: UKESM1: Description and Evaluation of the UK Earth System Model, *Journal of Advances in Modeling Earth Systems*, 11, 4513–4558, <https://doi.org/10.1029/2019ms001739>, 2019.
- 555 Sillman, S.: The use of NO<sub>y</sub>, H<sub>2</sub>O<sub>2</sub>, and HNO<sub>3</sub> as indicators for ozone-NO<sub>x</sub>-hydrocarbon sensitivity in urban locations, *Journal of Geophysical Research: Atmospheres*, 100, 14 175–14 188, <https://doi.org/10.1029/94JD02953>, 1995.
- Sillman, S.: The relation between ozone, NO<sub>x</sub> and hydrocarbons in urban and polluted rural environments, *Atmospheric Environment*, 33, 1821–1845, [https://doi.org/10.1016/S1352-2310\(98\)00345-8](https://doi.org/10.1016/S1352-2310(98)00345-8), 1999.
- Sillman, S. and West, J.: Reactive nitrogen in Mexico City and its relation to ozone-precursor sensitivity: results from photochemical models, 560 *Atmospheric Chemistry and Physics*, 9, 3477–3489, 2009.
- Silver, B., Reddington, C., Arnold, S., and Spracklen, D.: Substantial changes in air pollution across China during 2015–2017, *Environmental Research Letters*, 13, 114 012, <https://doi.org/10.1088/1748-9326/aae718>, 2018.
- Simon, H., Reff, A., Wells, B., Xing, J., and Frank, N.: Ozone trends across the United States over a period of decreasing NO<sub>x</sub> and VOC emissions, *Environmental science & technology*, 49, 186–195, 2015.
- 565 Stevenson, D., Dentener, F., Schultz, M., Ellingsen, K., Van Noije, T., Wild, O., Zeng, G., Amann, M., Atherton, C., and Bell, N.: Multimodel ensemble simulations of present-day and near-future tropospheric ozone, *Journal of Geophysical Research: Atmospheres*, 111, <https://doi.org/10.1029/2005JD006338>, 2006.
- Stevenson, D. S., Young, P. J., Naik, V., Lamarque, J. F., Shindell, D. T., Voulgarakis, A., Skeie, R. B., Dalsoren, S. B., Myhre, G., Berntsen, T. K., Folberth, G. A., Rumbold, S. T., Collins, W. J., MacKenzie, I. A., Doherty, R. M., Zeng, G., van Noije, T. P. C., Strunk, A., 570 Bergmann, D., Cameron-Smith, P., Plummer, D. A., Strode, S. A., Horowitz, L., Lee, Y. H., Szopa, S., Sudo, K., Nagashima, T., Josse, B.,

- Cionni, I., Righi, M., Eyring, V., Conley, A., Bowman, K. W., Wild, O., and Archibald, A.: Tropospheric ozone changes, radiative forcing and attribution to emissions in the Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP), *Atmos. Chem. Phys.*, 13, 3063–3085, <https://doi.org/10.5194/acp-13-3063-2013>, 2013.
- 575 Stock, Z. S., Russo, M. R., and Pyle, J. A.: Representing ozone extremes in European megacities: the importance of resolution in a global chemistry climate model, *Atmospheric Chemistry and Physics*, 14, 3899–3912, <https://doi.org/10.5194/acp-14-3899-2014>, 2014.
- Tarasick, D., Galbally, I. E., Cooper, O. R., Schultz, M. G., Ancellet, G., Leblanc, T., Wallington, T. J., Ziemke, J., Liu, X., Steinbacher, M., et al.: Tropospheric Ozone Assessment Report: Tropospheric ozone from 1877 to 2016, observed levels, trends and uncertainties, *Elementa: Science of the Anthropocene*, 7, 2019.
- 580 Thornhill, G., Collins, W. J., Kramer, R. J., Olivie, D., Skeie, R. B., O'Connor, F. M., Abraham, N. L., Checa-Garcia, R., Bauer, S. E., Deushi, M., Emmons, L. K., Forster, P. M., Horowitz, L. W., Johnson, B., Keeble, J., Lamarque, J. F., Michou, M., Mills, M. J., Mulcahy, J. P., Myhre, G., Nabat, P., Naik, V., Oshima, N., Schulz, M., Smith, C. J., Takemura, T., Tilmes, S., Wu, T., Zeng, G., and Zhang, J.: Effective radiative forcing from emissions of reactive gases and aerosols – a multi-model comparison, *Atmos. Chem. Phys.*, 21, 853–874, <https://doi.org/10.5194/acp-21-853-2021>, 2021a.
- 585 Thornhill, G., Collins, W., Olivie, D., Skeie, R. B., Archibald, A., Bauer, S., Checa-Garcia, R., Fiedler, S., Folberth, G., Gjermundsen, A., Horowitz, L., Lamarque, J. F., Michou, M., Mulcahy, J., Nabat, P., Naik, V., O'Connor, F. M., Paulot, F., Schulz, M., Scott, C. E., Séférian, R., Smith, C., Takemura, T., Tilmes, S., Tsigaridis, K., and Weber, J.: Climate-driven chemistry and aerosol feedbacks in CMIP6 Earth system models, *Atmos. Chem. Phys.*, 21, 1105–1126, <https://doi.org/10.5194/acp-21-1105-2021>, 2021b.
- 590 Turnock, S. T., Allen, R. J., Andrews, M., Bauer, S. E., Deushi, M., Emmons, L., Good, P., Horowitz, L., John, J. G., Michou, M., Nabat, P., Naik, V., Neubauer, D., O'Connor, F. M., Olivie, D., Oshima, N., Schulz, M., Sellar, A., Shim, S., Takemura, T., Tilmes, S., Tsigaridis, K., Wu, T. W., and Zhang, J.: Historical and future changes in air pollutants from CMIP6 models, *Atmospheric Chemistry and Physics*, 20, 14 547–14 579, <https://doi.org/10.5194/acp-20-14547-2020>, 2020.
- van Marle, M. J. E., Kloster, S., Magi, B. I., Marlon, J. R., Daniau, A. L., Field, R. D., Arneth, A., Forrest, M., Hantson, S., Kehrwald, N. M., Knorr, W., Lasslop, G., Li, F., Mangeon, S., Yue, C., Kaiser, J. W., and van der Werf, G. R.: Historic global biomass burning emissions for CMIP6 (BB4CMIP) based on merging satellite observations with proxies and fire models (1750-2015), *Geoscientific Model Development*, 595 10, 3329–3357, <https://doi.org/10.5194/gmd-10-3329-2017>, 2017.
- van Vuuren, D. P., Kriegler, E., O'Neill, B. C., Ebi, K. L., Riahi, K., Carter, T. R., Edmonds, J., Hallegatte, S., Kram, T., Mathur, R., and Winkler, H.: A new scenario framework for Climate Change Research: scenario matrix architecture, *Climatic Change*, 122, 373–386, <https://doi.org/10.1007/s10584-013-0906-1>, 2014.
- 600 Walters, D., Baran, A. J., Boutle, I., Brooks, M., Earnshaw, P., Edwards, J., Furtado, K., Hi, P., Lock, A., Manners, J., Morcrette, C., Mulcahy, J., Sanchez, C., Smith, C., Stratton, R., Tennant, W., Tomassini, L., Van Weverberg, K., Vosper, S., Willett, M., Browse, J., Bushell, A., Carslaw, K., Dalvi, M., Essery, R., Gedney, N., Hardiman, S., Johnson, B., Johnson, C., Jones, A., Jones, C., Mann, G., Milton, S., Rumbold, H., Sellar, A., Ujiie, M., Whitall, M., Williams, K., and Zerroukat, M.: The Met Office Unified Model Global Atmosphere 7.0/7.1 and JULES Global Land 7.0 configurations, *Geoscientific Model Development*, 12, 1909–1963, <https://doi.org/10.5194/gmd-12-1909-2019>, 2019.
- 605 Wang, T., Xue, L., Brimblecombe, P., Lam, Y. F., Li, L., and Zhang, L.: Ozone pollution in China: A review of concentrations, meteorological influences, chemical precursors, and effects, *Science of the Total Environment*, 575, 1582–1596, <https://doi.org/10.1016/j.scitotenv.2016.10.081>, 2017.

- Wang, W., van der A, R., Ding, J., van Weele, M., and Cheng, T.: Spatial and temporal changes of the ozone sensitivity in China based on satellite and ground-based observations, *Atmos. Chem. Phys.*, 21, 7253–7269, <https://doi.org/10.5194/acp-21-7253-2021>, 2021.
- 610 Whalley, L. K., Slater, E. J., Woodward-Massey, R., Ye, C., Lee, J. D., Squires, F., Hopkins, J. R., Dunmore, R. E., Shaw, M., Hamilton, J. F., et al.: Evaluating the sensitivity of radical chemistry and ozone formation to ambient VOCs and NO<sub>x</sub> in Beijing, *Atmospheric Chemistry and Physics*, 21, 2125–2147, 2021.
- Wild, O.: Modelling the global tropospheric ozone budget: exploring the variability in current models, *Atmospheric Chemistry and Physics*, 7, 2643–2660, <https://doi.org/10.5194/acp-7-2643-2007>, 2007.
- 615 Wild, O. and Prather, M. J.: Global tropospheric ozone modeling: Quantifying errors due to grid resolution, *Journal of Geophysical Research-Atmospheres*, 111, <https://doi.org/10.1029/2005jd006605>, 2006.
- Ye, L., Wang, X., Fan, S., Chen, W., Chang, M., Zhou, S., Wu, Z., and Fan, Q.: Photochemical indicators of ozone sensitivity: application in the Pearl River Delta, China, *Frontiers of Environmental Science & Engineering*, 10, 1–14, 2016.
- Young, P. J., Archibald, A. T., Bowman, K. W., Lamarque, J. F., Naik, V., Stevenson, D. S., Tilmes, S., Voulgarakis, A., Wild, O., Bergmann, D., Cameron-Smith, P., Cionni, I., Collins, W. J., Dalsoren, S. B., Doherty, R. M., Eyring, V., Faluvegi, G., Horowitz, L. W., Josse, B., Lee, Y. H., MacKenzie, I. A., Nagashima, T., Plummer, D. A., Righi, M., Rumbold, S. T., Skeie, R. B., Shindell, D. T., Strode, S. A., Sudo, K., Szopa, S., and Zeng, G.: Pre-industrial to end 21st century projections of tropospheric ozone from the Atmospheric Chemistry and Climate Model Intercomparison Project (ACCMIP), *Atmospheric Chemistry and Physics*, 13, 2063–2090, <https://doi.org/10.5194/acp-13-2063-2013>, 2013.
- 625 Young, P. J., Naik, V., Fiore, A. M., Gaudel, A., Guo, J., Lin, M., Neu, J., Parrish, D., Rieder, H., and Schnell, J.: Tropospheric Ozone Assessment Report: Assessment of global-scale model performance for global and regional ozone distributions, variability, and trends, *Elementa: Science of the Anthropocene*, 6, <https://doi.org/10.1525/elementa.265>, 2018.
- Zhang, J. F., Wei, Y. J., and Fang, Z. F.: Ozone Pollution: A Major Health Hazard Worldwide, *Frontiers in Immunology*, 10, <https://doi.org/10.3389/fimmu.2019.02518>, 2019.