



*Supplement of*

## **Large contributions of soil emissions to the atmospheric nitrogen budget and their impacts on air quality and temperature rise in North China**

**Tong Sha et al.**

*Correspondence to:* Tong Sha ([tong-sha@sust.edu.cn](mailto:tong-sha@sust.edu.cn)) and Jun Wang ([jun-wang-1@uiowa.edu](mailto:jun-wang-1@uiowa.edu))

The copyright of individual parts of the supplement might differ from the article licence.

## 1 **S1. Approach of converting the MEIC inventory to the model-ready formats.**

2 The MEIC inventory used in this study has a spatial resolution of  $0.25^\circ \times 0.25^\circ$   
3 and is constructed on an equal latitude-longitude grid. The model domain, however,  
4 employs a Lambert projection, which results in a misalignment between these two grid  
5 systems. We thus implement a spatial interpolation method to reallocate emission fluxes  
6 to the model grid. The descriptions are as follows:

7 Under the Lambert projection, the model grids are rectangular, while the MEIC  
8 grids are deformed, approximating a trapezoid shape. For calculating the emission of  
9 each model grid, we determine which MEIC grid may fall in that grid, based on their  
10 central latitude and longitude coordinates, and then apply the principle that the ratio of  
11 emissions is equivalent to the ratio of areas between the model grid and MEIC grid.  
12 The area of the MEIC grid is denoted as  $A$  and its corresponding emission is denoted  
13 as  $E$ , and the area of the model grid is denoted as  $a$ , thus the emission fluxes of model  
14 grid  $e$  are expressed as  $e = E \times a / A$ . However, if the spatial resolution of the simulation  
15 domain is coarser than MEIC, the model grid often falls within multiple MEIC grids  
16 and this method would have errors. We thus divide the coarser model grid into  $n \times n$   
17 finer subgrids. Since the spatial resolution of the nested domain is 9 km in this study,  
18 we choose  $n$  as 9. We apply the aforementioned method to calculate the emissions of  
19 each finer model subgrid, and then sum up the emissions of  $n \times n$  finer subgrids to  
20 obtain the total emissions of the model grid. This method ensures a more accurate  
21 allocation of MEIC emission to the model grid, despite different spatial resolutions of  
22 the simulation domain.

23

## 24 S2. Parameterization of HONO sources

25 In the present study, we incorporated five additional HONO sources in the WRF-  
26 Chem model, as described below.

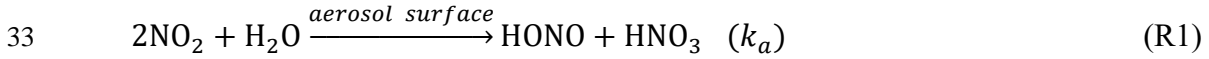
### 27 1. Direct traffic emissions

28 The traffic emission was calculated by a HONO/NO<sub>x</sub> ratio of 1.7% (Rappenglück  
29 et al., 2013), which is the same as the setting of (Zhang et al., 2019).

### 30 2. The HONO source from soil emissions

31 See section 2.1.2 in the manuscript.

### 32 3. Heterogeneous source on aerosol surface



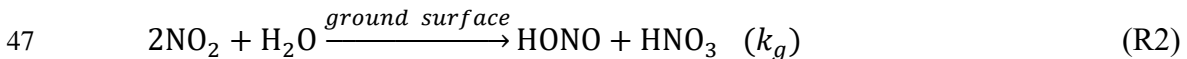
34 Most studies suggested that the heterogeneous reaction of NO<sub>2</sub> to HONO was first  
35 order in NO<sub>2</sub> (Finlayson-Pitts et al., 2003; Saliba et al., 2000), thus for the NO<sub>2</sub>  
36 heterogeneous reaction on the aerosol surface, the first-order reaction rate constant  $k_a$   
37 is estimated by (Li et al., 2010) and (Zhang et al., 2016) as follows:

$$38 \quad k_a = \frac{1}{4} \cdot v_{\text{NO}_2} \cdot \left(\frac{S_a}{V}\right) \cdot \gamma_{\text{a-NO}_2} \quad (\text{S1})$$

39 where  $v_{\text{NO}_2}$  is the mean molecular velocity of NO<sub>2</sub> (m s<sup>-1</sup>),  $S_a/V$  is the aerosol surface  
40 to volume ratio (m<sup>-1</sup>) representing the surface available for heterogeneous reaction.  
41  $\gamma_{\text{a-NO}_2}$  is the uptake coefficient of NO<sub>2</sub> at the aerosol surface, which was set to be  $1 \times$   
42  $10^{-6}$  for nighttime, and a higher value of  $2 \times 10^{-5}$  applied for daytime when the light  
43 intensity (LI) was lower than 400 W m<sup>-2</sup>, whereas we linearly scaled it with solar  
44 radiation when the light intensity was higher than 400 W m<sup>-2</sup> (equation 2).

$$45 \quad \gamma_{\text{a-NO}_2} = \begin{cases} 1 \times 10^{-6} \text{ (nighttime)} \\ 2 \times 10^{-5} \cdot \left(\frac{\text{LI}}{400}\right) \text{ (daytime, LI} \geq 400\text{W m}^{-2}\text{)} \\ 2 \times 10^{-5} \text{ (daytime, LI} < 400\text{W m}^{-2}\text{)} \end{cases} \quad (\text{S2})$$

### 46 4. Heterogeneous source on ground surface



48 For the NO<sub>2</sub> heterogeneous reaction on ground surface (R2), the first-order  
49 reaction rate constant  $k_g$  is estimated by (Zhang et al., 2016) as follows:

$$50 \quad k_g = \frac{1}{8} \cdot v_{\text{NO}_2} \cdot \left(\frac{S_g}{V}\right) \cdot \gamma_{\text{g-NO}_2} \quad (\text{3})$$

51 where  $v_{\text{NO}_2}$  is the mean molecular velocity of  $\text{NO}_2$  ( $\text{m s}^{-1}$ ),  $S_g/V$  represents the ground  
52 surface to volume ratio. Over the urban areas as defined by the MODIS land use data,  
53 we adopted a constant  $S_g/V$  value of  $0.3 \text{ m}^{-1}$ . For the vegetation-covered areas, the leaf  
54 area index (LAI,  $\text{m}^2/\text{m}^2$ ) and the height of the first model layer ( $H$ , m) were used to  
55 estimate the surface area to volume ratio following the method in (Sarwar et al., 2008):

$$56 \quad \frac{S_g}{V} = \frac{2 \times \text{LAI}}{H} \quad (4)$$

57  $\gamma_{\text{g-NO}_2}$  is the uptake coefficient of  $\text{NO}_2$  at the ground surface and is assumed to be the  
58 same as that for aerosol surface. The heterogeneous reaction of  $\text{NO}_2$  on the ground  
59 surface was only considered within the first model layer, whereas that on the aerosol  
60 surface was treated in all model layers.

#### 61 5. Inorganic nitrate photolysis

62 The photolysis reaction of particulate nitrate in the atmosphere to produce HONO  
63 and  $\text{NO}_2$  (R3) was added in the WRF-Chem model following the work of (Fu et al.,  
64 2019).



66 The photolysis rate of particulate nitrate was estimated by a  $J_{\text{nitrate}}/J_{\text{HNO}_3}$  ratio of  
67  $\frac{8.3 \times 10^{-5}}{7 \times 10^{-7}}$ , where  $J_{\text{HNO}_3}$  is the photolysis rate of gaseous  $\text{HNO}_3$  simulated online in the  
68 model.

69

70 **Table S1** Total annual N fertilizer application from 2006 to 2018 (unit: 10 Gg N yr<sup>-1</sup>),  
 71 and the adjustment coefficient (2006 vs. 2018, unit: %) for N fertilizer application in  
 72 each province.

Province	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015	2016	2017	2018	2006vs.2018
<b>Neimenggu</b>	64	68	73	80	80	81	83	89	97	99	98	95	86	34.5
<b>Gansu</b>	37	38	38	38	38	38	40	40	41	41	39	34	33	-9.6
<b>Ningxia</b>	16	17	17	17	18	18	18	19	18	17	17	17	16	4.5
<b>Shaanxi</b>	76	81	81	87	88	91	98	99	96	94	92	90	89	16.2
<b>Shanxi</b>	41	41	40	39	40	39	39	38	36	34	32	28	25	-38
<b>Hebei</b>	155	156	153	153	153	152	152	151	151	148	145	140	114	-26.2
<b>Beijing</b>	8	7	7	7	7	7	6	6	5	5	4	4	3	-61.3
<b>Tianjin</b>	12	13	12	12	12	11	11	11	11	10	9	7	6	-53.6
<b>Henan</b>	235	239	240	239	244	245	246	244	241	239	228	220	202	-14.3
<b>Shandong</b>	194	193	170	165	163	159	160	158	154	151	146	139	131	-32.5
<b>Jiangsu</b>	183	183	181	182	180	174	169	166	164	162	158	151	146	-20.4
<b>Anhui</b>	112	111	112	112	112	114	114	114	112	108	105	101	96	-14.4
<b>Huibe</b>	140	143	149	154	156	159	159	153	146	139	134	128	113	-19.5
<b>Chongqing</b>	46	48	50	50	49	50	50	50	50	50	48	47	46	-1.1
<b>Sichuan</b>	125	128	129	131	130	129	128	126	126	125	122	117	112	-10.1
<b>Liaoning</b>	63	65	66	67	68	70	68	70	68	66	60	57	55	-13.5

74 **Table S2.** Soil categories used in (Oswald et al., 2013).

<b>No.</b>	<b>Soil Category in Oswald et al. (2013)</b>
<b>S1</b>	eucalyptus forest, Grose Valley, Australia
<b>S2</b>	tropical rain forest, Suriname
<b>S3</b>	coniferous forest, Hohenpei ßenberg, Germany
<b>S4</b>	coniferous forest, Fichtelgebirge, Germany
<b>S5</b>	pasture, Hawkesbury River flood plain, Australia
<b>S6</b>	open woody savannah, Dahra, Senegal
<b>S7</b>	open woody savannah, Agoufou, Mali
<b>S8</b>	grassland, Mainz-Finthen, Germany
<b>S9</b>	pasture, Hohenpei ßenberg, Germany
<b>S10</b>	stone desert, Ruta B 376, Chile
<b>S11</b>	maize field, Grignon, France
<b>S12</b>	wheat field, Mainz-Finthen, Germany
<b>S13</b>	jujube field, Qiemo, China
<b>S14</b>	cotton field, Qiemo, China
<b>S15</b>	jujube field, Mingfeng, China
<b>S16</b>	stone desert, Sache, China
<b>S17</b>	cotton field, Milan, China

75

76 **Table S3.** Emission factor of 20 soil biomes based on MODIS land cover types.

<b>ID</b>	<b>MODIS land cover type</b>	<b>MEDIUM</b>	<b>HIGH</b>	<b>LOW</b>
1	Evergreen needleleaf forest	S3, S4	S3	S4
2	Evergreen broadleaf forest	S2	S2	S2
3	Deciduous needleleaf forest	S3, S4	S3	S4
4	Deciduous broadleaf forest	S1	S1	S1
5	Mixed forest	S1, S2, S3, S4	S2	S4
6	Closed shrublands	S6, S7, S8	S6	S8
7	Open shrublands	S6, S7	S6	S7
8	Woody savannas	--	--	--
9	Savannas	S6, S7	S6	S7
10	Grasslands	S8	S8	S8
11	Permanent wetlands	--	--	--
12	Croplands	S5, S9, S11, S12, S14, S17	S12	S9
13	Urban and built-up	--	--	--
14	Cropland/Natural vegetation mosaic	S8, S5, S9, S11, S12, S14, S17	S12	S9
15	Snow and ice	--	--	--
16	Barren or sparsely vegetated	S16, S10	S10	S16
17	water	--	--	--
18	Wooded tundra	--	--	--
19	Mixed tundra	--	--	--
20	Barren Tundra	--	--	--

77

**Table S4.** The optimum SHONO fluxes used in this study and other literature.

ID	MODIS land cover type	Optimum SHONO fluxes ( $\text{ng m}^{-2} \text{s}^{-1}$ )	References (land cover type in local scale)
1	Evergreen needleleaf forest	<b>0.549</b>	this study
2	Evergreen broadleaf forest	<b>2.872</b>	this study
3	Deciduous needleleaf forest	<b>0.549</b>	this study
4	Deciduous broadleaf forest	<b>0.887</b>	this study
		<b>1.214</b>	this study
		1.3	Zhou et al. (2011)
5	Mixed forest	0.01-104.72 (mean=16.45)	Wu et al. (2022)
		0.2-208 (mean=50)	Wang et al. (2023)
6	Closed shrublands	<b>20.57</b>	this study
7	Open shrublands	<b>29.779</b>	this study
8	Woody savannas	--	--
		<b>9.926</b>	this study
9	Savannas	1.1	Weber B (2015)
		<b>2.154</b>	this study
10	Grasslands	1.0	Twigg et al. (2011)
		0.1-74.27(mean =17.57)	Wu et al. (2022)
11	Permanent wetlands	--	--
		<b>30.036</b>	this study
		1.42-376.01(mean =119.8)	Wu et al. (2019, 2022)
		0.84±2.38	Meng et al. (2022)
12	Croplands	-1.32-7.69 (mean=2.94)	Tang et al. (2020)
		3.21	Xue et al. (2019)
		16-484	Wang et al. (2023)
13	Urban and built-up	--	--
14	Cropland/Natural vegetation mosaic	<b>25.847</b>	this study
15	Snow and ice	--	--
16	Barren or sparsely vegetated	<b>14.742</b>	this study



		1.5	Weber B (2015)
		5.38-288.23 (mean=57.06)	Wu et al. (2022)
<b>17</b>	water	--	--
<b>18</b>	Wooded tundra	--	--
<b>19</b>	Mixed tundra	--	--
<b>20</b>	Barren Tundra	--	--

---

79

80 **Table S5.** Contribution of soil NO<sub>x</sub> and HONO emissions to monthly average surface  
 81 concentrations of NO<sub>2</sub> and HONO (unit: %) in July 2018.

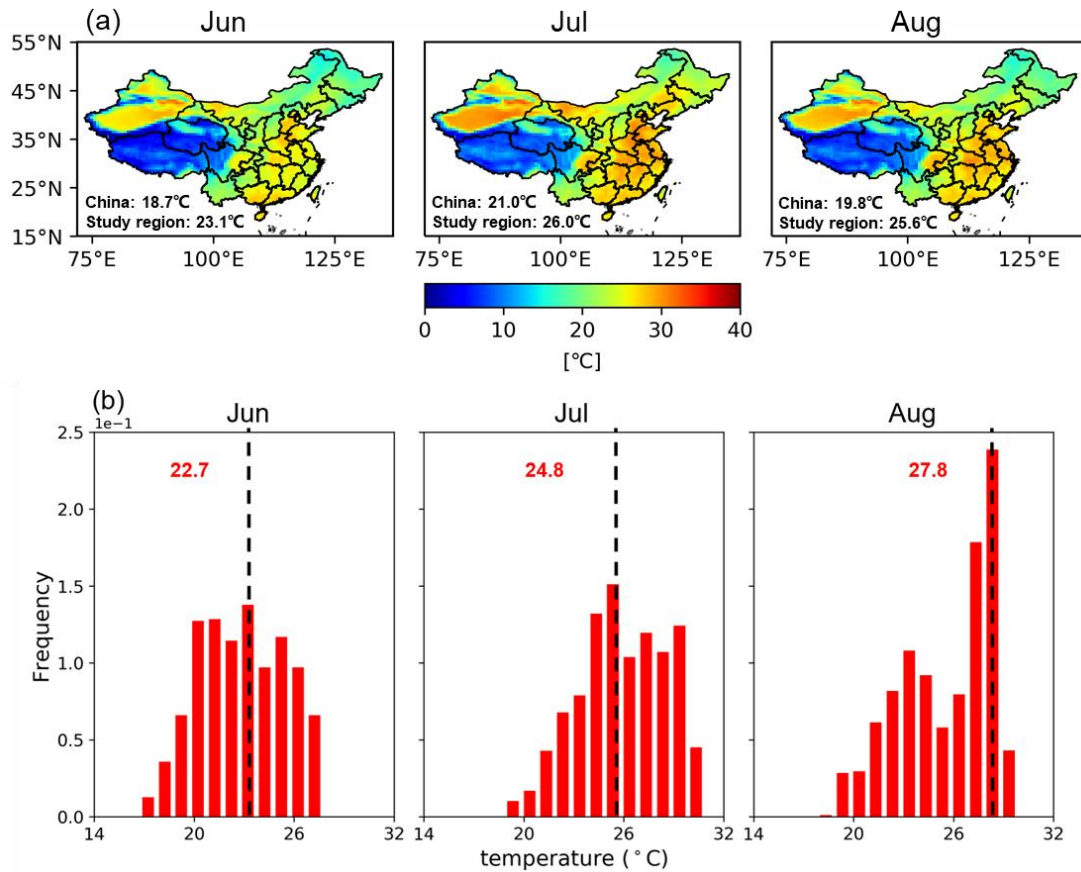
Contribution	Surface NO <sub>2</sub> concentrations			Surface HONO concentrations		
	Study region (CL)	BTH(CL)	FWP(CL)	Study region (CL)	BTH(CL)	FWP(CL)
SNO <sub>x</sub>	30.3(33.2)	37.1(39.5)	31.8(38.6)	6.2(5.7)	7.8(7.6)	4.95(4.2)
SHONO	3.1(2.3)	1.8(1.75)	2.7(3.1)	35.6 (38.7)	36.7(38.6)	38.0(42.7)
Soil Nr	32.7(34.7)	38.4(40.5)	33.9(41.3)	38.2(20.0)	40.3(42.0)	40.1(44.6)

82

83 **Table S6.** Effect of soil NO<sub>x</sub> and HONO emissions on monthly average surface  
 84 concentrations of MDA8 O<sub>3</sub>, max-1h OH, and nitrate in BTH and FWP region during  
 85 July 2018 (unit: %).

Change	MDA8 O <sub>3</sub>			max-1h ·OH			nitrate		
	Study region (CL)	BTH (CL)	FWP (CL)	Study region (CL)	BTH (CL)	FWP (CL)	Study region (CL)	BTH (CL)	FWP (CL)
Soil NO <sub>x</sub>	15.3 (17.4)	13.9 (15.0)	14.6 (15.6)	-31.3 (-21.6)	-28.4 (-13.5)	-38.6 (-24.8)	17.8 (22.4)	29.6 (41.3)	27.6 (32.8)
Soil HONO	3.3 (3.0)	3.5 (3.8)	2.8 (3.1)	10.0 (13.4)	9.3 (13.1)	10.3 (17.5)	10.4 (11.3)	10.9 (14.2)	13.5 (15.2)
Soil Nr	18.2 (20.0)	16.9 (18.1)	17.2 (18.6)	-24.3 (-12.5)	-22.6 (-4.4)	-32.2 (-13.6)	31.8 (35.8)	42.4 (57.8)	42.7 (49.9)

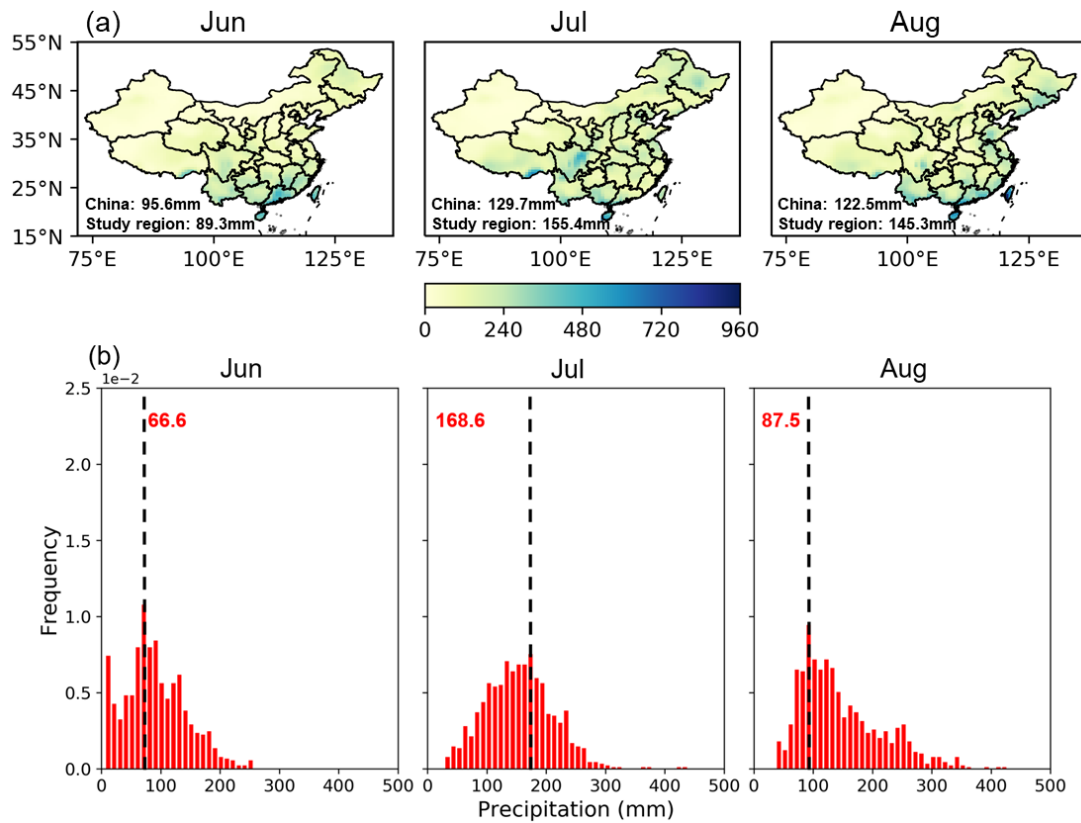
86



87

88 **Figure S1.** (a) Distribution of monthly average air temperature at 2m (T2) from the  
 89 MERRA-2 dataset during June-August 2018. The statistics over lower left corner are  
 90 the monthly average T2 over China and the study region. (b) Frequency of the monthly  
 91 average T2 in the study region during June-August 2018. The statistics on each panel  
 92 are the T2 corresponding to the highest frequency.

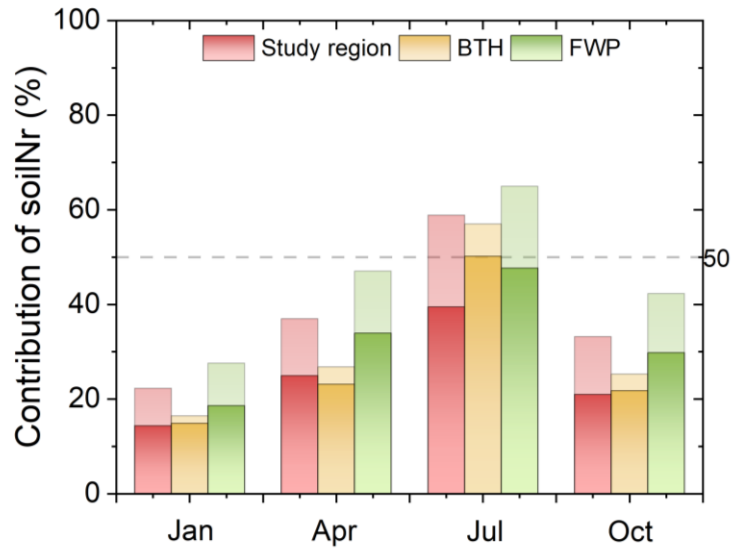
93



94

95 **Figure S2.** Distribution of monthly total precipitation from the MERRA-2 dataset  
 96 during June-August 2018. The statistics over lower left corner are the total monthly  
 97 precipitation over China and the study region. (b) Frequency of the monthly total  
 98 precipitation in the study region during June-August 2018. The statistics on each panel  
 99 are the total precipitation amount corresponding to the highest frequency.

100

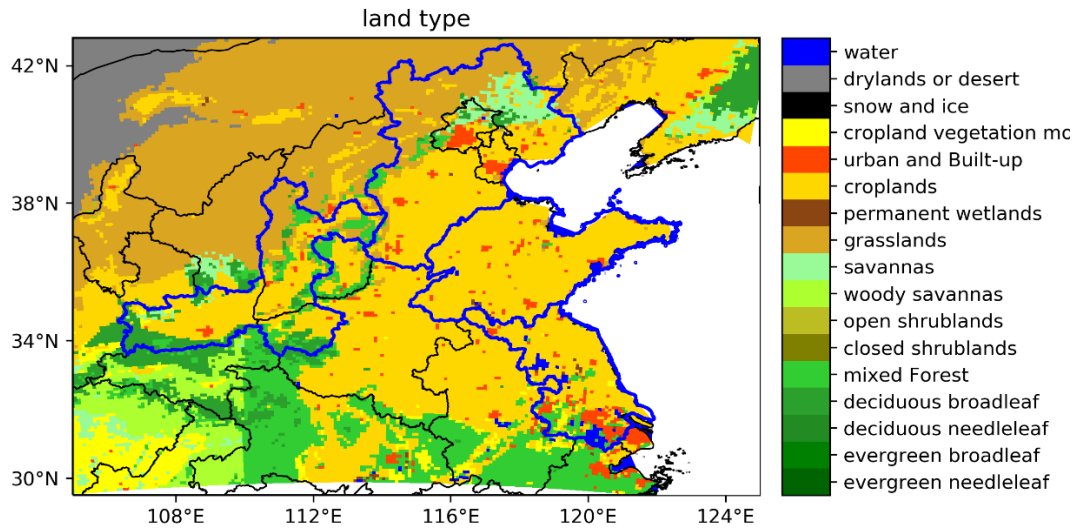


101

102 **Figure S3.** Monthly proportion of soil Nr emissions to anthropogenic NO<sub>x</sub> emissions  
 103 during January, April, July, and October in the study region, BTH, and FWP regions.

104 The darker columns with borders are statistics for the whole region, while the lighter  
 105 columns are statistics for croplands. The gray horizontal dotted line in the figure  
 106 represents a 50% proportion.

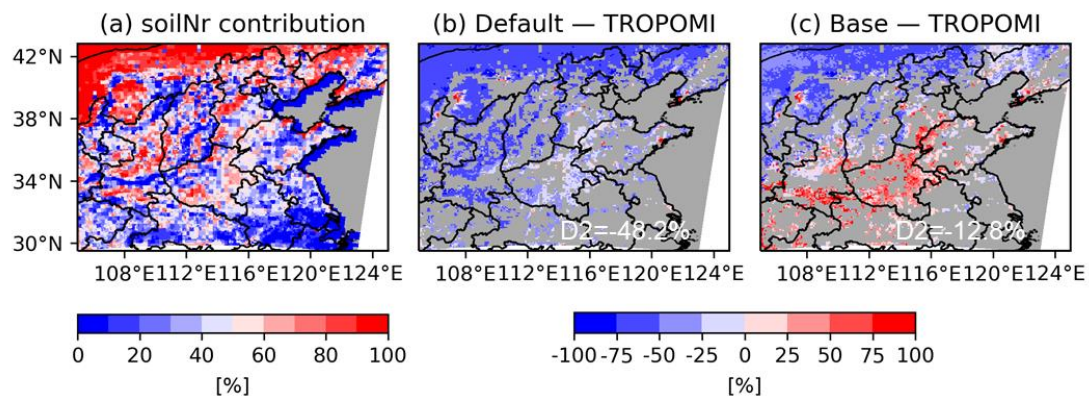
107



108

109 **Figure S4.** The land cover type over the simulation domain.

110

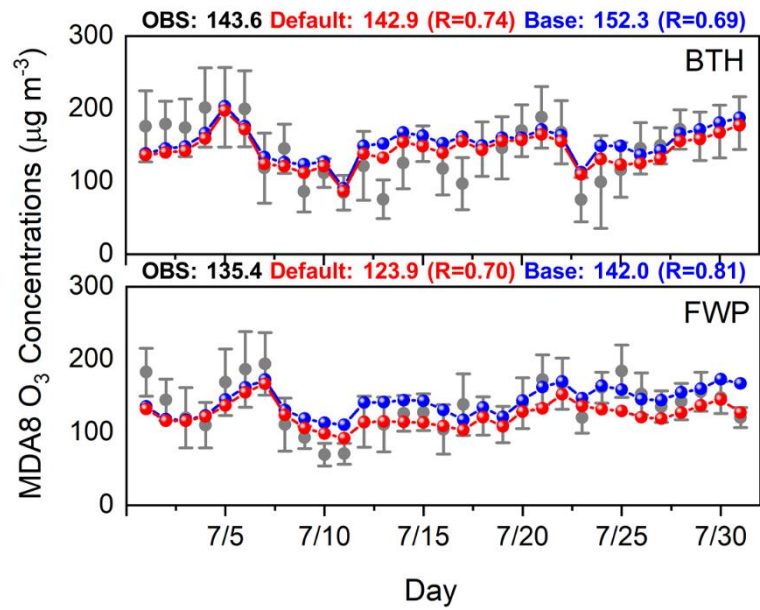


111

112 **Figure S5.** (a) Distribution of simulated contribution of soil Nr emissions to total Nr  
 113 emissions, which includes the sources from anthropogenic emissions, soil emissions,  
 114 and biomass burning. The difference of monthly mean tropospheric NO<sub>2</sub> VCD from  
 115 TROPOMI observations and simulations ((b) Default, (c) Base). The grids where soil  
 116 Nr emissions contributions are lower than 50% are masked to better compare the  
 117 difference of observations and simulations in rural areas surrounding cities. Statistics  
 118 in each panel are the mean value averaged over the study region in July 2018.

119

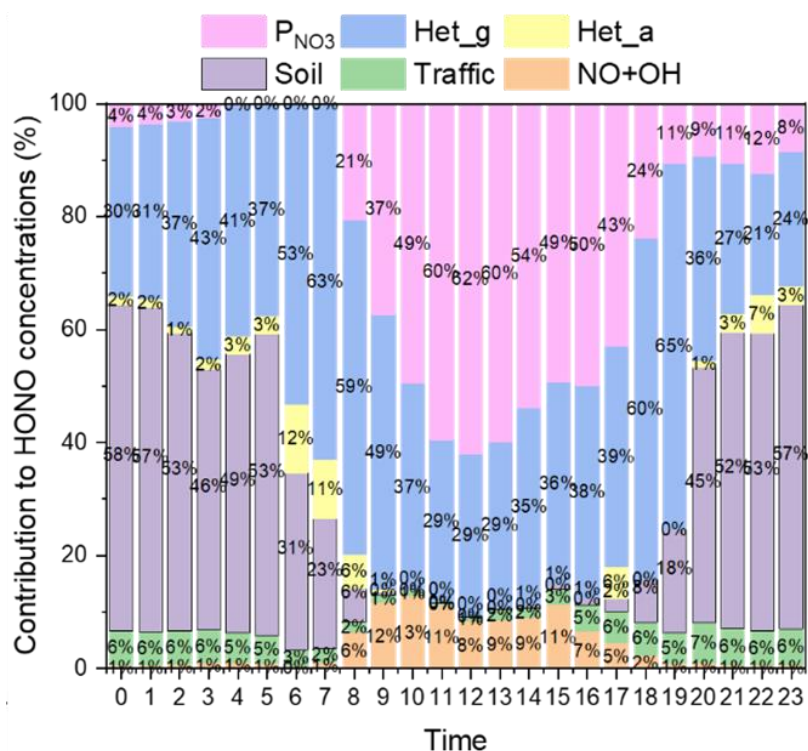




120

121 **Figure S6.** Time series of observed (grey circles with bars representing the standard  
 122 deviations) and simulated (Default in red and Base in blue) surface MDA8 O<sub>3</sub>  
 123 concentrations in the BTH and FWP regions in July 2018, with the mean value and  
 124 temporal correlation coefficients (R) shown in the upper corner.

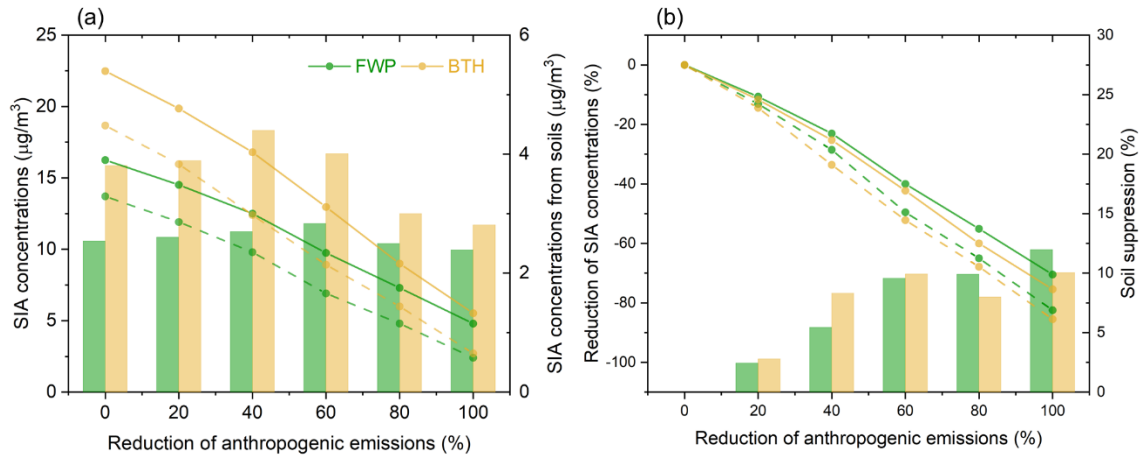
125



126

127 **Figure S7.** Average diurnal variations of contributions of different HONO sources to  
 128 the simulated surface HONO at a rural station in Nanjing during July 2018. (P<sub>NO3</sub>,  
 129 Het\_g, Het\_a, Soil, Traffic, and NO+OH represent HONO sources from the inorganic  
 130 nitrate photolysis in the atmosphere, NO<sub>2</sub> heterogeneous reactions on ground and  
 131 aerosol surfaces, soil emissions, traffic emissions, and the gas-phase formation,  
 132 respectively).

133



134

135 **Figure S8.** The responses of secondary inorganic aerosols (SIA) concentrations to the  
 136 reductions of anthropogenic emissions (taking into account the SO<sub>2</sub>, NO<sub>x</sub>, primary  
 137 PM<sub>2.5</sub>, VOCs, and CO reduced by 20%, 40%, 60%, 80%, and 100%) relative to July  
 138 2018 levels in the presence (solid line) and absence (dotted line) of soil Nr emissions.  
 139 (The lines in panels (a-b) are SIA concentrations and the relative reductions in SIA  
 140 concentrations under different anthropogenic emission reductions, respectively. The  
 141 bars (right y-axis) in panel (a) show the corresponding SIA concentrations from soil Nr  
 142 emissions (denoted as SIA concentrations from soils) under different anthropogenic  
 143 emission reductions, which are determined as the difference between the solid and  
 144 dotted lines. The bars (right y-axis) in panel (b) show the suppression of SIA reduction  
 145 due to the existence of soil Nr emissions (denoted as soil suppression), which are  
 146 determined as the difference between the solid and dotted lines. Green lines and bars  
 147 are the results in the FWP region, and the yellow are the results in the BTH region.)

148

149 **Reference**

- 150 Finlayson-Pitts, B. J., Wingen, L. M., Sumner, A. L., Syomin, D., and Ramazan, K. A.:  
151 The heterogeneous hydrolysis of NO<sub>2</sub> in laboratory systems and in outdoor and  
152 indoor atmospheres: An integrated mechanism, *Phys. Chem. Chem. Phys.*, 5, 223-  
153 242, 10.1039/b208564j, 2003.
- 154 Fu, X., Wang, T., Zhang, L., Li, Q., Wang, Z., Xia, M., Yun, H., Wang, W., Yu, C., Yue,  
155 D., Zhou, Y., Zheng, J., and Han, R.: The significant contribution of HONO to  
156 secondary pollutants during a severe winter pollution event in southern China,  
157 *Atmos. Chem. Phys.*, 19, 1-14, 10.5194/acp-19-1-2019, 2019.
- 158 Li, G., Lei, W., Zavala, M., Volkamer, R., Dusanter, S., Stevens, P., and Molina, L. T.:  
159 Impacts of HONO sources on the photochemistry in Mexico City during the  
160 MCMA-2006/MILAGO Campaign, *Atmos. Chem. Phys.*, 10, 6551-6567,  
161 10.5194/acp-10-6551-2010, 2010.
- 162 Meng, F., Qin, M., Fang, W., Duan, J., Tang, K., Zhang, H., Shao, D., Liao, Z., Feng,  
163 Y., Huang, Y., Ni, T., Xie, P., Liu, J., and Liu, W.: Measurement of HONO flux  
164 using the aerodynamic gradient method over an agricultural field in the Huaihe  
165 River Basin, China, *J. Environ. Sci.*, 114, 297-307, 10.1016/j.jes.2021.09.005,  
166 2022.
- 167 Oswald, R., Behrendt, T., Ermel, M., Wu, D., Su, H., Cheng, Y., Breuninger, C.,  
168 Moravek, A., Mougín, E., Delon, C., Loubet, B., Pommerening-Röser, A., Sörgel,  
169 M., Pöschl, U., Hoffmann, T., Andreae, M. O., Meixner, F. X., and Trebs, I.:  
170 HONO Emissions from Soil Bacteria as a Major Source of Atmospheric Reactive  
171 Nitrogen, *Science*, 341, 1233-1235, 10.1126/science.1242266, 2013.
- 172 Rappenglück, B., Lubertino, G., Alvarez, S., Golovko, J., Czader, B., and Ackermann,  
173 L.: Radical precursors and related species from traffic as observed and modeled at  
174 an urban highway junction, *J. Air Waste Manage. Assoc.*, 63, 1270-1286, 2013.
- 175 Saliba, N. A., Mochida, M., and Finlayson-Pitts, B. J.: Laboratory studies of sources of  
176 HONO in polluted urban atmospheres, *Geophys. Res. Lett.*, 27, 3229-3232,  
177 10.1029/2000gl011724, 2000.
- 178 Sarwar, G., Roselle, S. J., Mathur, R., Appel, W., Dennis, R. L., and Vogel, B.: A

179 comparison of CMAQ HONO predictions with observations from the Northeast  
180 Oxidant and Particle Study, *Atmos. Environ.*, 42, 5760-5770,  
181 10.1016/j.atmosenv.2007.12.065, 2008.

182 Tang, K., Qin, M., Fang, W., Duan, J., Meng, F., Ye, K., Zhang, H., Xie, P., Liu, J., Liu,  
183 W., Feng, Y., Huang, Y., and Ni, T.: An automated dynamic chamber system for  
184 exchange flux measurement of reactive nitrogen oxides (HONO and NO<sub>x</sub>) in  
185 farmland ecosystems of the Huaihe River Basin, China, *Sci. Total Environ.*, 745,  
186 140867, 10.1016/j.scitotenv.2020.140867, 2020.

187 Twigg, M. M., House, E., Thomas, R., Whitehead, J., Phillips, G. J., Famulari, D.,  
188 Fowler, D., Gallagher, M. W., Cape, J. N., Sutton, M. A., and Nemitz, E.:  
189 Surface/atmosphere exchange and chemical interactions of reactive nitrogen  
190 compounds above a manured grassland, *Agric. For. Meteorol.*, 151, 1488-1503,  
191 10.1016/j.agrformet.2011.06.005, 2011.

192 Wang, Y., Fu, X., Wang, T., Ma, J., Gao, H., Wang, X., and Pu, W.: Large Contribution  
193 of Nitrous Acid to Soil-Emitted Reactive Oxidized Nitrogen and Its Effect on Air  
194 Quality, *Environ. Sci. Technol.*, 57, 3516-3526, 10.1021/acs.est.2c07793, 2023.

195 Weber B, W. D., Tamm A, et al. : Biological soil crusts accelerate the nitrogen cycle  
196 through large NO and HONO emissions in drylands, *Proc. Natl. Acad. Sci. U.S.A.*,  
197 112(50): 15384-15389., 2015.

198 Wu, D., Horn, M. A., Behrendt, T., Müller, S., Li, J., Cole, J. A., Xie, B., Ju, X., Li, G.,  
199 Ermel, M., Oswald, R., Fröhlich-Nowoisky, J., Hoor, P., Hu, C., Liu, M., Andreae,  
200 M. O., Pöschl, U., Cheng, Y., Su, H., Trebs, I., Weber, B., and Sörgel, M.: Soil  
201 HONO emissions at high moisture content are driven by microbial nitrate  
202 reduction to nitrite: tackling the HONO puzzle, *ISME J.*, 13, 1688-1699,  
203 10.1038/s41396-019-0379-y, 2019.

204 Wu, D., Zhang, J., Wang, M., An, J., Wang, R., Haider, H., Xu-Ri, Huang, Y., Zhang,  
205 Q., Zhou, F., Tian, H., Zhang, X., Deng, L., Pan, Y., Chen, X., Yu, Y., Hu, C., Wang,  
206 R., Song, Y., Gao, Z., Wang, Y., Hou, L., and Liu, M.: Global and Regional Patterns  
207 of Soil Nitrous Acid Emissions and Their Acceleration of Rural Photochemical  
208 Reactions, *J. Geophys. Res.: Atmos.*, 127, 10.1029/2021jd036379, 2022.

209 Xue, C., Ye, C., Zhang, Y., Ma, Z., Liu, P., Zhang, C., Zhao, X., Liu, J., and Mu, Y.:  
210 Development and application of a twin open-top chambers method to measure soil  
211 HONO emission in the North China Plain, *Sci. Total Environ.*, 659, 621-631,  
212 10.1016/j.scitotenv.2018.12.245, 2019.

213 Zhang, J., An, J., Qu, Y., Liu, X., and Chen, Y.: Impacts of potential HONO sources on  
214 the concentrations of oxidants and secondary organic aerosols in the Beijing-  
215 Tianjin-Hebei region of China, *Sci. Total Environ.*, 647, 836-852,  
216 10.1016/j.scitotenv.2018.08.030, 2019.

217 Zhang, L., Wang, T., Zhang, Q., Zheng, J., Xu, Z., and Lv, M.: Potential sources of  
218 nitrous acid (HONO) and their impacts on ozone: A WRF-Chem study in a  
219 polluted subtropical region, *J. Geophys. Res.: Atmos.*, 121, 3645-3662,  
220 10.1002/2015jd024468, 2016.

221 Zhou, X., Zhang, N., TerAvest, M., Tang, D., Hou, J., Bertman, S., Alaghmand, M.,  
222 Shepson, P. B., Carroll, M. A., Griffith, S., Dusanter, S., and Stevens, P. S.: Nitric  
223 acid photolysis on forest canopy surface as a source for tropospheric nitrous acid,  
224 *Nat. Geosci.*, 4, 440-443, 10.1038/ngeo1164, 2011.

225